

Title: Synthesizing theories of natural resource management and governance

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Integration of social-ecological knowledge from different disciplines is lacking

Without such integration, previous work is under-leveraged and progress is stymied

We present the results of a collaborative effort to achieve such integration

We do so by formalizing theories of natural resource governance from multiple fields

We conclude by discussing these theories and their relationships with each other

1. Introduction

Understanding how human-environment interactions can generate more sustainable outcomes remains one of the key challenges for contemporary scientific inquiry (Clark 2007; DeFries et al. 2012). For this, the integration of knowledge from different disciplines has been recognized as key (Newell et al. 2005; Ostrom 2007). However, despite calls for a unified “sustainability science” (Clark 2007), the analysis of human-environment interactions remains dispersed across a wide range of fields, each with their own set of theories (VanWey et al. 2005; Armitage 2007; Turner and Robbins 2008). These fields include the study of social-ecological systems, coupled-natural-human systems, resilience, common-pool resources, environmental, resource and ecological economics, environmental anthropology, psychology and sociology, political ecology, sustainability science, conservation biology, and land-use and land cover change, among others. Collectively, these fields have produced an impressive amount of knowledge regarding the factors affecting human behavior and subsequent environmental outcomes.

To date, however, no formal effort has been made to consistently express the theories developed within these disciplines in a way that could both document the body of accumulated theory and specify testable hypotheses. Theories are a critical way of representing scientific knowledge by expressing important causal relationships among a set of related concepts. In the absence of such an effort, several problems persist.

First, to the extent that progress in sustainability science (or any science) should, at least in part, be measured by the accumulation of empirically supported theoretical statements, progress in the study of human-environment interactions cannot be easily evaluated and understood. Relatedly, meaningful discussions regarding the level of evidentiary support for various theories in the relevant fields are needed, but are difficult to have in the absence of a repository of such theories.

Second, empirical work cannot be as effective or efficient without a thorough understanding of the theories that could be used to deductively generate relevant hypotheses for a particular study area. Because of the diversity of disciplinary perspectives represented in human-environment studies, as well as the lack of cross-disciplinary engagements between different approaches, many scholars are simply unaware of alternative hypotheses or theoretical perspectives from other fields. What researchers measure and analyze in a particular case, then, may not reflect the full body of knowledge within the literature. Indeed, in much empirical work on human-environment interactions, even in articles that employ a standardized framework such as the one introduced by Ostrom (2007), the common practice is to under-explain the (presumably theoretical) reasons for which variables examined were chosen (Thiel et al. 2015). As a result, researchers may continue to collect facts and document patterns inductively, without testing theories generated by their colleagues, and the science of sustainability may perpetually remain in an exploratory phase.

An under-emphasis on theoretical consolidation has been a challenge to the practice of science for some time. Forscher (1963, 339) remarked on this problem in his brickmakers parable: “The brickmakers became obsessed with the making of bricks. When reminded that the ultimate goal was edifices, not bricks, they replied that, if enough bricks were available, the builders would be able to select what was necessary and still continue to construct edifices...and so it happened that the land became flooded with bricks.” For the study of human-environment interactions, this is a practical problem as well as an academic one. A synthesis of knowledge on human-environment interactions will enable policy analysts and policymakers to more effectively diagnose particular problems and bring to bear the relevant knowledge to explore possible solutions.

51 The purpose of this paper is to present the results of a collaborative effort to consistently document
52 theories from several fields focused on human-environment interactions. This work is inspired by,
53 but ultimately quite distinct from, a framework for the study of social-ecological systems developed
54 by Ostrom (2007, 2009). Our research is part of a larger project known as the Social-Ecological
55 Systems Meta-Analysis Database (SESMAD) project (Cox 2014). Within this database, we document
56 many important sustainability-oriented theories generated by scholars across multiple disciplines
57 and perspectives.

58

59 **2. Materials and Methods**

60

61 **2.1. Preliminary work**

62

63 Prior to coding theories into the SESMAD database, the SESMAD team members met multiple times
64 at in-person workshops to establish the overall approach and a common understanding of the main
65 concepts and protocols to be used. In order to facilitate the storage of the data needed to express
66 theories and variables, we constructed a relational database, with a table for each necessary object:
67 theories, variables, and studies, as well as several others less relevant for this paper. Data entry
68 forms for each object were also constructed to guide data entry.

69

70 The structure of the database reflects several previous synthetic, comparative efforts, especially the
71 International Forestry Resources and Institutions (IFRI) Project (Gibson et al. 2000; Poteete and
72 Ostrom 2004) and Ostrom's (2007) SES framework. This previous work established protocols for
73 exploring human impacts on the environment as interactions between an environmental resource,
74 one or more actor groups, and an overarching governance system. Starting with this
75 conceptualization, we collaborated with a team of database experts to produce the digital
76 infrastructure needed to consistently code information on interactions between humans and their
77 environment. The data are currently stored on a server as an Oracle database and are publicly
78 accessible via the SESMAD website (<https://sesmad.dartmouth.edu>).

79

80 To support the coding of the SESMAD theories, we populated the database with all of the variables
81 that the group agreed to be relevant for the analysis of social-ecological systems. This was in fact
82 done to support the coding of empirical cases of social-ecological systems, which the database also
83 supports. Each group member was asked to come up with a list of the most important concepts
84 from their field and draft an entry in the database for a corresponding variable. This coding
85 included identifying what the primary coder perceived to be the most prominent studies relevant to
86 the meaning of this variable. During several in-person meetings, these drafts were then reviewed in
87 sub-groups to ensure that the variables were described according to their most widely understood
88 meaning and usage. During the theory coding process new variables were added and reviewed as
89 necessary for particular theories.

90

91 **2.2. Sampling and coding steps**

92

93 To bound our inquiry we developed the following formal criteria for inclusion in the SESMAD
94 database. To be included, each theory had to (1) be prevalent in at least one relevant scientific field,
95 (2) include both social and ecological variables, (3) point to a relatively clear causal chain between
96 one or more independent variables and an outcome for a natural resource (e.g., a forest or a
97 fishery), and (4) be described in at least one peer-reviewed study.

98

99 We were unable to sample theories in the traditional sense because there is no bounded sampling
100 frame or population from which to select, which is part of the problem we are attempting to

101 address. Thus instead of trying to include all environmental resource governance theories, we
 102 focused on an important subset of governance theories which deal primarily with rational choice
 103 models of human decision making and the role of collective action and compliance in affecting
 104 environmental outcomes. Most of these theories were developed from studies of relatively small to
 105 regional-scale systems, and were the core theoretical approaches that underlay Ostrom's SES
 106 framework (Ostrom 2007), although they have recently also been applied to larger-scale systems
 107 (Cox 2014; Fleischman et al. 2014).

108
 109 Adopting this approach meant leaving out several sets of theories, including those that do not focus
 110 on local-to-regional contexts or do not rely on rational choice models of human decision making,
 111 but are still widely used to study environmental governance. Thus, we did not include theories
 112 developed by international relations scholars to study international environmental regimes (Young
 113 2011). We also do not include behavioral theories that focus on norms and perceptions, as opposed
 114 to economic incentives and collective decision-making, like the theory of planned behavior
 115 (Fishbein and Ajzen 2010). Additionally, we did not include many theories used in political ecology,
 116 which often focus on factors that are difficult to operationalize as we have in the database, or seek
 117 to explain how important categories come to be socially constructed, and how power mediates this
 118 process (Goldman et al. 2011). Finally, following criterion number three above meant that we did
 119 not include theories that involved recursive relationships involving mutual feedback between two
 120 or more variables. This led theories such as the adaptive cycle (Gunderson and Holling 2002) and
 121 co-evolutionary arguments to be excluded.

122
 123 The first step in the sampling process was to have the group brainstorm the most relevant theories
 124 in the disciplines they knew best. This created an initial list of theories from multiple disciplines.
 125 For each theory, a member of the team entered a first draft of the theory into the shared database
 126 using the standard data entry forms. This member was the "expert" for this theory, and was
 127 primarily responsible for its coding. For each theory so created, the expert for each theory led a
 128 discussion amongst members of the larger group. If the larger team decided that a theory could be
 129 effectively formalized by entry into the SESMAD database, a second draft was produced that was
 130 then reviewed one final time by the lead author to ensure that it complied with the coding
 131 protocols, to which we now turn.

132 133 **2.3. Coding protocols**

134
 135 To aid in and standardize theory formalization, we developed a basic coding protocol that was
 136 implemented in a set of online data entry forms. The most important element of this protocol is its
 137 definition of a "theory." This term does not have one universal meaning across (or sometimes
 138 within) scientific disciplines. Our own working definition of a theory is based on the discussion of
 139 theories and two closely related concepts, frameworks and models, in McGinnis and Ostrom
 140 (2014):

141
 142 "A framework provides the basic vocabulary of concepts and terms that may be used to construct
 143 the kinds of causal explanations expected of a theory. Frameworks organize diagnostic, descriptive,
 144 and prescriptive inquiry. A theory posits specific causal relationships among core variables. In
 145 contrast, a model constitutes a more detailed manifestation of a general theoretical explanation in
 146 terms of the functional relationships among independent and dependent variables important in a
 147 particular setting. Just as different models can be used to represent different aspects of a given
 148 theory, different theoretical explanations can be built upon the foundation of a common conceptual
 149 framework."

150

151 Frameworks then populate the scientist's world with a set of conceptual objects and (non-causal)
 152 relationships among them. Theories are statements that stipulate causal interactions among these
 153 objects, and models formalize theories mathematically, generally for the purposes of testing them.
 154 Based on this and similar work, we have adopted the following definition of a theory, or theoretical
 155 statement:

157 **Theory:** A statement that describes (1) a relationship between an outcome and a set of
 158 independent variables, the values of which are argued to be *sufficient* for predicting the outcome,
 159 and (2) a mechanism by which this relationship occurs.

161 We interpret theories to express sufficient conditions rather than necessary conditions. For
 162 example, the theory of the Tragedy of the Open-Access Commons argues that open-access
 163 conditions and low property security are sufficient for environmental deterioration. This would
 164 predict that for every case of open-access and low tenure security, the condition of the commons
 165 will suffer. It does not mean that every case of environmental degradation involves an open-access
 166 property regime and low tenure security (which would be a necessary relationship). To the extent
 167 that the logical inverse of a theory can be meaningfully expressed, this complementary theory
 168 would express a necessary relationship: some type of property regime with high property security
 169 is necessary for environmental sustainability.

171 This definition closely ties theories to another important scientific object: variables, and we have
 172 adopted the following working definition of variables:

174 **Variable:** A scientifically important concept that has been assigned a level of measurement and
 175 range of possible values.

177 A variable defines an attribute of a particular object. The objects that our variables describe are
 178 components of a social-ecological system, roughly based on Ostrom's social-ecological framework.
 179 The three components that SESMAD variables can describe are (1) actor groups, (2) governance
 180 systems, and (3) environmental commons.

182 Variables can take on multiple roles in a theory (Table 1): dependent variables are the outcomes
 183 that a theory explains, and independent variables do the explaining. An independent variable can be
 184 specified as an underlying or proximate variable, depending on its immediacy to the final outcome
 185 (Geist and Lambin 2002). For theories with multiple independent variables, such variables are
 186 essentially INUS conditions: see Regin's (2000) discussion of INUS conditions originally developed
 187 by Mackie (1965).

189 Outcomes can similarly be specified as intermediate or final. Each theory must have at least one
 190 final outcome that describes the predicted result for a natural resource commons, and an
 191 intermediate outcome is specified if there is an outcome of interest (e.g. collective action from the
 192 commons literature) that is seen as facilitating the final outcome. Finally, a third type of
 193 independent variable that is available is a moderating variable. A moderating variable influences
 194 the strength of a relationship between two other variables in a theory.

196 Any variable that contributes to a theory is assigned one of these roles as well as a qualitative
 197 explanation of its role. It is also assigned a value, which can be either (1) one or more values from
 198 its range of allowable values, or (2) a descriptor of a change in its value. This way of formalizing
 199 theories is consistent with the hypothetico-deductive approach to scientific practice, whereby
 200 theories are developed and enable predictions that are then tested via the deductive specification of

201 their observational implications (what we would expect to see in the world if they were true). The
 202 value of each variable in a SESMAD theory is one such observational implication that can be used to
 203 test the theory in an empirical setting (King et al. 1994).

204
 205 Each SESMAD theory has an open text field which summarizes the causal narrative that links each
 206 of the variables together. We also considered the ways in which theories could be simple vs.
 207 complex based on the variables they contain. A simple theory specifies a relationship between one
 208 independent variable and an outcome, where this relationship is not seen to be contingent on any
 209 third moderating variable (no interaction effects). We recorded multiple ways in which theories
 210 could depart from this simplicity. First, we recorded how many steps they involved (with the
 211 maximum being four for theories that involve underlying and proximate independent variables and
 212 intermediate and final outcomes). Theories with four steps were informally labeled as “process”
 213 theories. A process theory, such as Gilded Traps (Steneck et al. 2011) describes a sequence of
 214 related events that culminate in one or more related outcomes. A theory can also gain complexity
 215 by containing many variables at a single step. Such “list” theories list out many facilitating
 216 conditions needed in order to produce a particular outcome. Finally, complex theories are
 217 characterized by interaction effects created by moderating variables.

218
 219 In addition to connecting to variables, theories can be connected to published studies (articles,
 220 books). This association can record whether the study empirically supports or contradicts the
 221 theory with which it is being associated, or simply describes it. At least one such study was
 222 recorded for each SESMAD theory.

223
 224 Finally, theories can have one of three types of associations to each other: nested, related and
 225 contradictory (Table 2). Table 2 shows three types of relationships. These relationships vary in the
 226 extent to which the relationships involved are symmetric, and in the extent to which they confer
 227 transitive relationships among theories. The only asymmetric relationship type is nested, in that A
 228 containing B does not imply that B contains A.

229
 230 The nested relationship is fully transitive: if A contains B and B contains C, then theory A contains
 231 theory C and all of its variables. Moreover, because theory A contains all of the variables (with the
 232 same, or when appropriate, opposite, values) as theory B, this also means that theory A inherits all
 233 of the other relationships held by theory B. These inherited relationships are implicit in the
 234 variables that theory A contains and are not formally coded for theory A. For example, if theory A
 235 contains theory B and theory B contains theory C and is contradictory with theory D, we did not
 236 code that theory A had any formal relationship with theories C or D.

237
 238 Related relationships are partially transitive. If theory A is related to theory B and theory B is
 239 related to theory C, whether or not theory A is then related to theory C depends on whether this
 240 relationship also satisfies the conditions for the related association. Theories that are related to the
 241 same theory are more likely to be related to each other in this way, but are not always related.
 242 Contradictory relationships are not transitive at all. However, if A and B are both contradictory to C,
 243 then it is likely that they are related to each other.

244 245 **2.4. Analysis**

246
 247 To analyze the coded theories, we first statistically summarized the results (Table 3), and then
 248 summarized the results by the scientific fields from which the theories were derived (Table 4).
 249 Following this, we examined the relationships between theories and variables (Figure 1) and with
 250 each other. To examine the theory-to-theory network, we used the Excel extension NodeXL. This

251 produced visualizations of the theory-to-theory network (Figures 3, 4, 5) as well as distributions of
 252 the number of connections between theories by type of connection (Figure 6). We also examined
 253 the extent to which theories tended to associate more or less with theories from their own or other
 254 scientific fields (Table 5). Finally, we conducted a basic correlational analysis and an exploratory
 255 cluster analysis to examine the extent to which different coded attributes of the theories tend to
 256 create different classes of theories (Table 6).

257 258 **3. Results**

259 Drawing upon 117 published studies, we inductively coded 63 theories using 81 variables (Table
 260 S1). All of the theories and variables are publicly viewable at the SESMAD website.

261 262 263 **3.1. Scientific fields and variables**

264 Table 3 defines the numerical attributes that were coded or calculated for each theory and provides
 265 summary statistics for each. Table 4 summarizes the results for each of the fields from which the
 266 theories were taken (see Appendix A for a full description of the fields). It also provides a brief
 267 description of each of these fields as they pertain to the environmental sector. The field with the
 268 most theories was Collective Action and the Commons, followed by a closely related field, Political
 269 Economy. Because our sample of cases cannot be said to be representative of a broader, ill-defined
 270 population of theories, we note that the distribution of theories across the different fields does not
 271 reflect the extent of theoretical content of different fields. We list it here merely to characterize our
 272 own sample of theories.
 273

274 Seven theories were recoded as Interdisciplinary. This somewhat understates the highly
 275 interdisciplinary nature of the majority of the theories that were coded. Also many theories could
 276 have been formally associated with an additional field. Those coded as interdisciplinary were coded
 277 as such because they could not be coded as belonging to one field over all others among the experts
 278 of the respective fields.
 279

280 While there was a fair amount of diversity in the variables used by the theories, as reflected in the
 281 distribution of the number of theories using each variable, several variables were used much more
 282 than others (Figure 1). The most used variable was Commons Condition Trend, which was used by
 283 59 theories. This variable was the final outcome for all theories with the exception of those from the
 284 resilience literature, which instead use Basin Switch as the final outcome variable. The second most
 285 common variable was Collective Action. This was used mostly as an intermediate outcome to
 286 explain environmental outcomes in the Commons literature (see Table 4). Next, the variable
 287 Compliance is used as an intermediate outcome in several other literatures, and also as a
 288 moderating variable (i.e., in theories that focused on the importance of policy instruments). The
 289 main difference between Compliance and Collective Action as intermediate outcomes is subtle, yet
 290 important. Theories emphasizing Compliance tend to view actors as the recipients of
 291 environmental policies imposed by external actors; while those emphasizing Collective Action
 292 highlight the role of those actors in the design and enforcement of environmental policies. The
 293 other important intermediate outcome variables were the level of Social-Ecological Fit between
 294 institutions and ecological systems and the Resilience of Ecosystems, which was unique to the
 295 resilience literature.
 296

297 Finally, the fields differed in the primary independent variable they used. Policy Instrument and
 298 Centralization were used in many theories because they were the primary independent variables in
 299 one or more fields. Centralization is used by many of the Political Economy theories, while Policy
 300

301 Instrument is used mostly by the Environmental and Resource Economics and Conservation
 302 Biology fields (Table 4).

303

304 3.2. Network results

305

306 Figures 2, 3 and 4 contain diagrams that depict all of the theories and their relationships to each
 307 other, using abbreviations of the theory names. Figure 5 provides a legend for each of the other
 308 figures. Figure 2 shows the theories involved in “related” and “contradictory” relationships, while
 309 Figure 3 shows the theories involved in “nested” relationship. Figure 4 shows all theories and all
 310 relationships. This figure shows a network with 63 nodes and 137 connections, with 56 being
 311 “related”, 59 being “nested” and 22 being “contradictory.” The connections between the theories
 312 are color-coded, as are the theories themselves based on the literature they were drawn from.

313

314 Figure 4 shows some degree of clustering by scientific field as well as the presence of theoretical
 315 “hubs”, or highly connected nodes. This first feature is unpacked further in table 5, which shows the
 316 counts of connections between theories of different fields. Overall 59 out of 137 (43%) connections
 317 in the full network (Figure 4) were between theories of the same field. The second feature is shown
 318 in figure 6, which contains the degree distributions for the full network and each of the three link
 319 types. Like many networks, the full network is right-skewed with a small number of theories having
 320 dis-proportionally large numbers of connections to other theories. The figure also shows that this
 321 skew comes primarily from the “nested” connections, and secondarily from the “related”
 322 connections, with the “contradictory” distribution being only slightly skewed.

323

324 The two largest network clusters are associated with the Commons and Political Economy fields
 325 (Figure 4). The main hub in the network, the theory Community Based Natural Resource
 326 Management (CBNRM), connects these two clusters. The clustering of the Political Economy
 327 theories reflects that most of these theories include the Centralization variable and make either
 328 similar or contradictory arguments with it. Closely related to this cluster are several theories from
 329 the Conservation Biology literature. The Commons cluster, meanwhile, has several connections to
 330 the Resilience theories at the bottom of the diagram. The final cluster (of four theories) using these
 331 links is located in the upper right, containing theories that argue about the role that technology
 332 plays in affecting environmental outcomes.

333

334 The theories from different fields differ in the extent to which they use the different types of theory-
 335 to-theory relationships (Figures 2, 3, 4, 5 Table 4). The Commons field relies almost exclusively on
 336 nested relationships, creating three highly connected hubs. One of these, Collective Action and the
 337 Commons, is used by all of the theories that assume a connection between collective action and
 338 positive commons outcomes. The other two hubs in this literature, Community Based Natural
 339 Resource Management (CBNRM) and CBNRM Design Principles, use many of the theories that
 340 attempt to explain collective action outcomes in mostly local contexts. The Resilience theories are
 341 similar in this regard, with one main hub (Conditions for General Resilience) containing the
 342 majority of the other resilience theories, as well as several from the Commons literature. The other
 343 two theories that were used by several others were Enforcement and Social-Ecological Fit Theory.
 344 The former is used as a pre-condition for success of theories that argue for the desirability of policy
 345 regimes (such as Individual Transferable Quotas), while Social-Ecological Fit (as a variable
 346 described in the theory of a similar name) is an intermediary outcome for many of the theories
 347 advocating for decentralized governance regimes.

348

349

350

351 **3.3. Correlational results**

352

353 Table 6 shows a basic correlation matrix between all of the attributes contained in Table 3. Here we
 354 see several things that further unpack results just presented. Least surprisingly, larger theories
 355 (theories with more variables) tend to have more theories nested in them, and partially as a result
 356 are more highly connected and more complex as measured by the final three dimensions. Theories
 357 that have “related” connections also tend to have “contradictory connections and to not be nested
 358 within other theories. Finally, list theories tend to have multiple steps as well, but tend not to have
 359 moderators. A cluster analysis was also conducted that showed two basic groups of theories: a
 360 small group of larger, more complex theories, and a large group of simpler ones. It is important to
 361 note that these correlational results are disproportionately driven by the large theories which
 362 appear as outliers in Figure 1.

363

364 **4. Discussion**

365

366 **4.1. Important elements and controversies**

367

368 As the results show, some variables and theories are much more widely used than others. The full
 369 network degree distribution is widely skewed, largely as a result of some theories containing or
 370 being contained by many other theories. It’s worth reiterating that the “contradictory” distribution
 371 does not indicate that there are particularly controversial theories, but this masks the importance
 372 of a particular variable that is behind the majority of these connections: the Centralization variable.
 373 The results indeed show that this variable is one of the more controversial in terms of the
 374 theoretical predictions made with it. At the same time, this analysis does not necessarily reveal
 375 other controversies within the literature by not recording situations of equifinality as
 376 contradictions. For example, many different arguments are made with the Policy Instrument
 377 variable, and controversies regarding what are the most effective instruments are widespread. This
 378 is not something the SESMAD database is currently well equipped to detect, however.

379

380 **4.2. Empirical utility of the database**

381

382 There are several ways in which researchers can use this database to support their empirical work.
 383 First, the SESMAD theory database can help researchers identify the most important variables to
 384 measure and theories to test for a particular empirical context or set of research questions.
 385 Through an iterative process, a researcher can combine their knowledge of a particular social-
 386 ecological system with the theories database to develop a list of variables that they want to
 387 measure to test the theories that are most relevant for their system.

388

389 For example, if a researcher is examining a case in which a group of local resource users are nested
 390 within a set of large-scale governance arrangements, s/he could examine the Multiple Levels
 391 variable and see what theories it is involved in. Each of these theories contains additional variables,
 392 and the value predicted for each of these variables is an observational implication that can be tested
 393 empirically (King et al. 1994). This functionality will only increase as additional theories are added
 394 to the database, with some being made specific to particular environmental sectors such as
 395 fisheries. The database currently supports this function, but it has not yet been implemented for
 396 more than a few theories, such as Forest Transition Theory. This function has been more
 397 thoroughly implemented for variables, with a subset of variables describing attributes and
 398 outcomes specific to marine protected areas (MPAs). These variables have not yet been used to
 399 formalize MPA-specific theories.

400

401 Additionally, the researcher can use the theory-to-theory relationships to explore what types of
 402 theories (and their associated variables) might be most relevant. For example, if a researcher is
 403 interested in testing the theory of Failure of Centralized Control in a particular case or set of cases,
 404 he or she could examine other theories that are associated with this theory (such as Centralization
 405 and Corruption, or Technical Solutions and Shifting the Burden and Gilded Traps) as they might
 406 also be relevant.

407
 408 Finally, because the SESMAD theories link to similarly formalized variables, which themselves are
 409 explicitly and systematically documented, researchers can use the database to help them decide
 410 how they want to measure the variables in the most relevant theories. Currently such details are
 411 poorly specified in much of the empirical social-ecological literature (Thiel et al. 2015).

412 413 **4.3. Testing theories**

414
 415 Whereas the reviewed literatures appear to have little difficulty in generating new theories, strong
 416 empirical tests of many of these theories are lacking beyond the cases in which they were
 417 developed. This likely results from several factors. First, many scholars may simply not be aware of
 418 many of the relevant theories they could test in their work. Second, it is analytically difficult to
 419 empirically test the complex theories that have come to dominate the literature on natural resource
 420 management and governance. Simple theories, which propose a direct relationship between a
 421 primary cause and an outcome, have been well-studied and enjoy considerable empirical support
 422 from both case-based analysis and large-n statistical studies (Gibson et al. 2005; Chhatre and
 423 Agrawal 2008; 2009; Cox et al. 2010; Persha et al. 2011; Villamayor-Tomas 2014). As represented
 424 in the total number of variables they used (Figure 1), the majority of the theories we coded were
 425 fairly simple, with the median number of variables used being four (Table 3).

426
 427 In contrast, there is far less broad, quantitative empirical support for more complex theories (such
 428 as Gilded Traps or the Failure of Centralized Control) that use many variables and which now
 429 dominate much of the discourse on sustainability. While the historical absence of formalization
 430 certainly impedes this goal, so has the absence of suitable methods to interrogate a large body of
 431 theory that is difficult to test (Poteete et al. 2010).

432
 433 The distinction between process vs. list-oriented theories has implications for empirical evaluation.
 434 List theories, such as CBNRM Design Principles, Polycentric Co-management and Conditions for
 435 General Resilience, are those which combine a set of simple theories into a list which generally
 436 predict success on the basis of adherence to that list. List theories do not necessarily contain
 437 interaction effects, but the presence of many variables as say proximate causes can imply a level of
 438 dependence among these variables in their effects on an outcome. For instance, Ostrom (1990, 90-
 439 91) speculates that something similar to her design principles might ultimately constitute a set of
 440 necessary or sufficient principles for robust community-based natural resource management; or
 441 that some combination of the design principles increases the likelihood of success. However,
 442 because there are thousands of possible combinations of the eight original design principles, it is
 443 effectively impossible to apply conventional statistical methods to test which combinations of these
 444 factors are likely to contribute to success. Other types of analysis suggest that the exact
 445 combination of sufficient conditions in a list theory may not matter, only that a minimum number of
 446 factors are present (Gutierrez et al. 2011). Nevertheless, the default approach to analyzing complex
 447 list theories is to effectively treat them as sets of simple theories, which are then tested individually
 448 (Quinn et al. 2007; Cox et al. 2010).

449

450 A number of theories involved many (up to four) steps. These process-oriented theories are those
451 that propose the mechanism by which a set of initial conditions affect changes in social-ecological
452 systems. The most prominent (or at least most complicated) examples of these in the database are
453 the Failure of Centralized Control and Gilded Traps theories, each of which details a fairly intricate
454 process by which their causal narratives unfold. The methodological implications for process-based
455 theories are somewhat different than list theories, in that they cannot as easily be decomposed into
456 semi-autonomous sub-theories. There are both qualitative (process-tracing) and quantitative
457 (structural and simultaneous equation modeling, mediation analysis) approaches to examining
458 these kinds of dynamics (Young et al. 2006), although to date these are not common methodologies
459 in the literatures we examined.

460
461 Finally, there are interaction-based theories, which suggest that the effects of variables are
462 contingent upon the state or magnitude of other factors. For instance, the Decentralization and
463 Local Capacity theory (Garces-Restrepo and Muñoz 2007) suggests that decentralization efforts are
464 unlikely to contribute to sustainable use of natural resources unless groups have the capacity to
465 undertake management responsibilities. In other words, the effects of decentralization are
466 contingent upon the presence of a group with high levels of social capital. Although interaction-type
467 theories are easier to test than list or process theories, and indeed several recent studies have done
468 so (Andersson and Agrawal 2011; Coleman 2011), they present their own analytical challenges.
469 First, models which include interaction terms will always have higher levels of multicollinearity
470 than models which do not, which reduces the power of statistical tests. Second, in a field where
471 scholars are often limited by small samples sizes (Poteete et al. 2010), the likelihood of developing
472 a strong empirical test involving more than two factors is unlikely.

473
474 Two final comments on this issue are warranted. First, our distinction between simple vs. complex
475 theories raises questions about the direction of theory development in sustainability science.
476 Simple theories explain highly general relationships that make sense only in probabilistic terms,
477 while complex theories are more specific about the conditions under which one can observe certain
478 outcomes, and thus are more deterministic in their predictions. The trade-off between developing
479 simple vs. complex theories resonates with that between the accuracy and meaning of predictions
480 (Cox 2008). Simple theories prioritize predictability (i.e., across different contexts), while complex
481 theories value meaning (i.e., to the contexts to which theories apply). Theory testing can advance
482 the field of sustainability science in both directions. Simple theories that are tested in new types of
483 contexts can result in new complex theories, while variables that fulfill similar roles in two or more
484 complex theories can inspire new simple theories.

485
486 Second, our analysis emerges from an effort to formalize theories; we have not systematically
487 looked at how the theories are used by scholars and policy makers. This distinction is not trivial, as
488 theories, particularly popular ones involving particular governance arrangements (primarily via
489 the Policy Instrument and Centralization variables), run the risk of being applied as “panacea”
490 prescriptions across diverse contexts (Ostrom 2007), or as blueprints against to which evaluate the
491 supposed sustainability of human-environment interactions. Anecdotally, we would say that some
492 list theories are becoming increasingly popular among scholars and policy makers. An example is
493 Polycentric Co-management. The ideas this theory expresses come from both a segment of the
494 Commons and Political Economy fields, which have discussed the benefits of polycentricity, and the
495 Resilience literature, where cross-scale adaptive co-management arrangements are frequently
496 lauded (Chaffin et al. 2014). Given its widespread support despite the challenges in testing this
497 theory across contexts, we suggest, with some hesitation given that this is a controversial argument,
498 that this theory increasingly resembles a “panacea”, as this is how we see its arguments playing out
499 in the literature. Several other theories we reviewed, such as Community-Based Natural Resource

500 Management (CBNRM) and Individual Transferable Quotas, have faced similar speculation
501 (Bromley 2009; Dressler et al. 2010).

502

503 **4.4. What counts as evidence?**

504

505 Testing complex theories is not only a tricky analytical issue, but a tricky philosophical one.
506 Scientists must consider what counts as evidence for or against such theories. The fact that complex
507 theories support the generation of multiple observational implications and empirical hypotheses is
508 a double-edged sword. On the positive side, it allows for more Popperian-like tests of theory, which,
509 as Cox (2015) explains, is the very condition that can enable strong small-n, case-based causal
510 inference:

511

512 “If the properties of a case are highly congruent with the expectations generated for it by a certain
513 theory, and there are no alternative theories with which the case is also congruent, we may
514 conclude that the results of the case are very unlikely except if they are explained by the theory, and
515 this can be seen as strong evidence in support of the theory. As George and Bennett (2005, 117) put
516 it: ‘an explanation of a case is more convincing if it is more unique, or if the outcome it predicts
517 could not have been expected from the best rival theory available.’ This is made possible by having
518 well-established theories with many independent predictions for a case. The more predictions, the
519 more points of congruence can potentially be established and the less likely it is that a confirmatory
520 case could be explained without the theory in question. ‘This process relies on Bayesian logic—the
521 more unique and unexpected the new evidence, the greater its corroborative power’ (George and
522 Bennett 2005, 219).”

523

524 On the other hand, it is unclear what constitutes a critical test for a theory. If a case affirms five out
525 of six observational implications (hypotheses) for a theory, is this confirmatory or contradictory? A
526 statistical perspective would probably dictate that it is more confirmatory than not. And many
527 scientists will feel comfortable in many circumstances to simply adjust a theory in light of new
528 evidence. On the face of it this is reasonable and acceptable practice. But the line between this and
529 simply making ad hoc changes to a cherished theory is fuzzy at best, and increasing analytical and
530 technological sophistication will not resolve this issue.

531

532 **5. Conclusion: limitations and future directions**

533

534 **5.1. Limitations**

535

536 There are several limitations to the approach we have taken. The first of these is the non-exhaustive
537 “sampling” process we used to produce our list of theories. The distribution of the number of
538 theories across the fields reflects the experiential bias of the researchers involved in this project,
539 rather than the actual extent of theoretical knowledge accumulated in these fields. As we are not
540 trying to generalize our results from our sample in the traditional way, this is not exactly an
541 inferential problem, but it is an important limitation nonetheless. Our intention is for our
542 formalization of theories to serve as a starting point to enable other researchers to add theories as
543 well.

544

545 Second, the way in which we have formally expressed the theories is not the only way of expressing
546 them. Particularly for the more complex theories, there is no standard understanding of their
547 arguments. Even for simple theories, we were sometimes challenged to arrive at a common
548 understanding of a rigorous formal expression. Relatedly, the three types of relationships between

549 theories presented in this study do not necessarily capture the full universe of potential relations
550 amongst theories.

551
552 Third, while this project benefited from the infrastructure of an online relational database
553 management system, other tools are possibly superior in recording scientific knowledge.
554 Technologies associated with the semantic web such as the Resource Description Framework (RDF)
555 and Web Ontology (OWL) ontology language, for example, are explicitly designed to store fields of
556 knowledge, including scientific knowledge (Frey and Cox 2015). A disadvantage of those
557 technologies is that they tend to be less well developed and have much less technical support and
558 available expertise than those associated with the dominant relational database paradigm.

559 **5.2. Future work**

560
561
562 The SESMAD database is not a final product. Rather, additional theories, variables, and connections
563 will be added to it over time. This process will only increase the need for a complementary process
564 to be undertaken, one which would document the extent of empirical support for the theories that
565 have been coded. This is particularly salient for pairs of theories that are coded as being mutually
566 contradictory. One promising approach is a type of vote-counting-based meta-analysis of empirical
567 analyses to document the extent to which different theories are supported or refuted. Without the
568 database it is difficult to see how this could be done systematically. Nevertheless, even with the
569 database to support such an effort, several challenges must be addressed.

570
571 The first of these is how to defensibly sample from the highly diffuse literatures that contribute to
572 these theories in a way that is representative of a hypothetical population of studies. Whereas in
573 this paper we were less concerned with obtaining a technically representative sample, to compare
574 different theories based on the amount of support in the literature would require such a sample of
575 hundreds to thousands of studies.

576
577 A second challenge is how to validly and reliably measure (and thus weight) the “level of empirical
578 support” offered for theories by different types of studies (e.g. case studies vs. large-n statistical
579 analyses) conducted at different levels of analysis and with different measurements of important
580 outcomes. This is a particularly challenging question for more complex theories as previous
581 discussed. This is also not purely a technical question, but rather one that strikes at the heart of
582 multiple fields and the philosophy of science behind many of our activities as researchers. It is a
583 challenge for the conduct of social-ecological science as a whole, for without addressing it sooner or
584 later, theoretical progress will be stymied as theories are essentially allowed to persist in a cloud of
585 incomplete mutual understanding among the relevant research communities.

586
587 There are several ways in which this challenge can be addressed. First, it could become a standard
588 practice in future empirical work to report the extent to which such work supports or contradicts
589 certain SESMAD theories. This would require a community-level agreement on how this
590 relationship is recorded, but if this were possible, it could greatly ease the burden on subsequent
591 efforts to code for such a relationship between empirics and theory after the fact. Second, and
592 relatedly, we can try to increase the comparability of empirical measurements, particularly on
593 dependent variables, in future work. The SESMAD theories as we have coded them reflect a
594 standard practice in much of the commons literature to conflate multiple ways of measuring
595 outcomes for environmental commons and the communities that depend on them. One way
596 forward here will involve the introduction of new variables to further specify what exactly is meant
597 by Common Condition Trend as the primary outcome used in SESMAD theories. Our group is

598 already engaged in such efforts with respect to outcomes specific to forest and fisheries
599 governance.

600

601 To conclude, research in sustainability science will continually be faced with the challenge of
602 synthesizing and comparing numerous case studies and theories. We hope our research encourages
603 the social-ecological research community to value the importance of reliability and consistency in
604 the measurement and expression of the concepts and theoretical statements they use to guide their
605 work.

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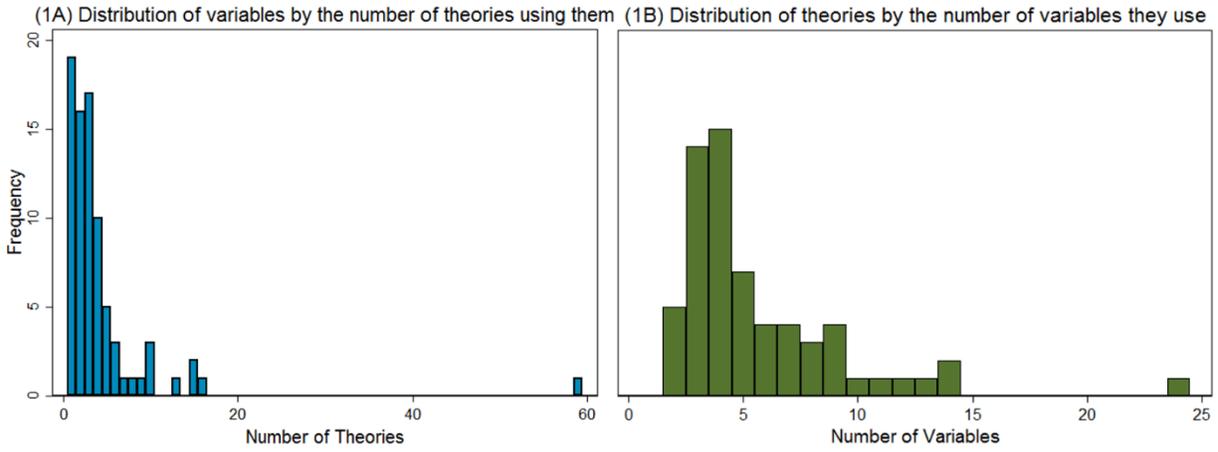
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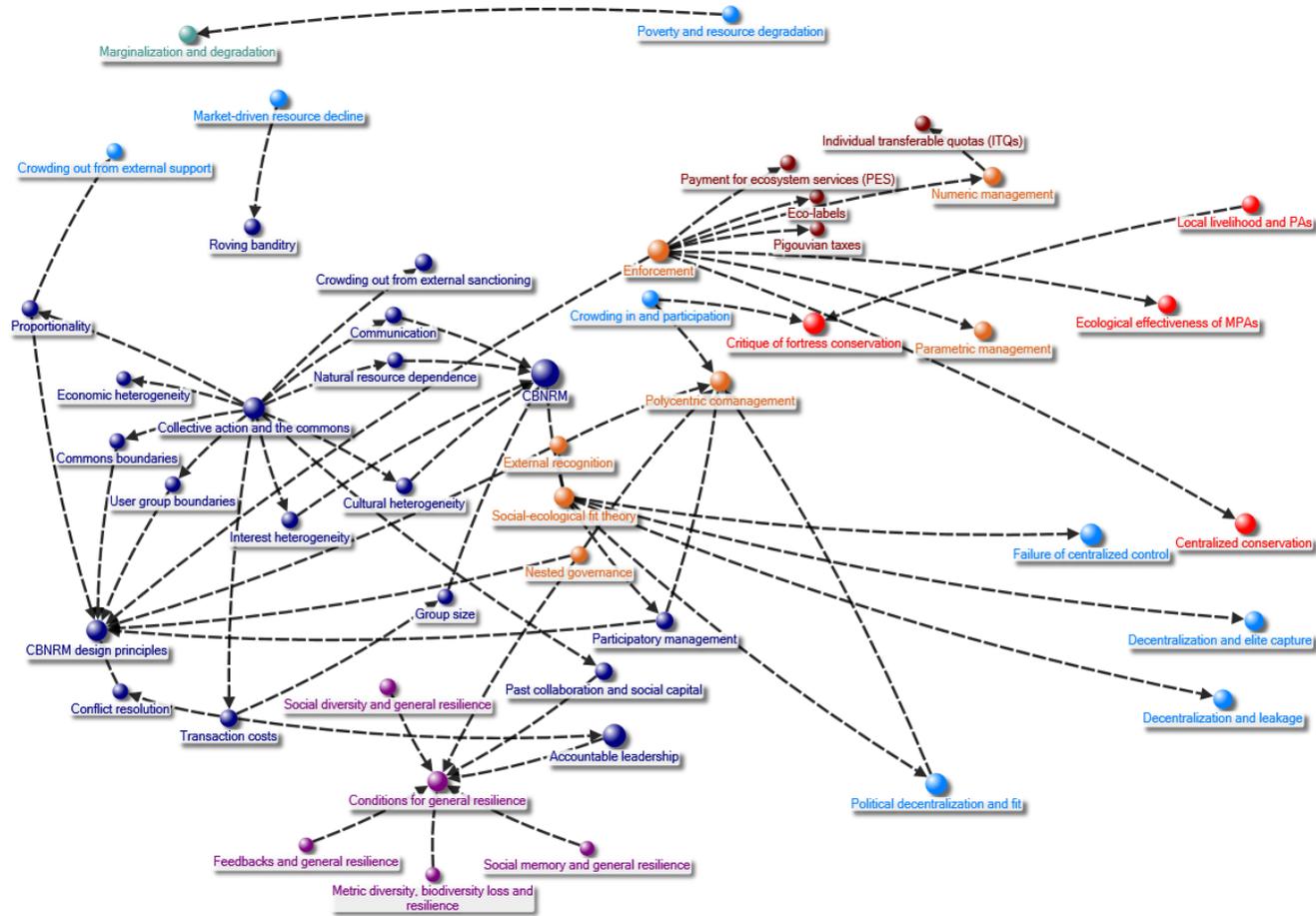


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Figure 1: Relationships between variables and theories

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Figure 3: Network diagram of nested theory-to-theory connections

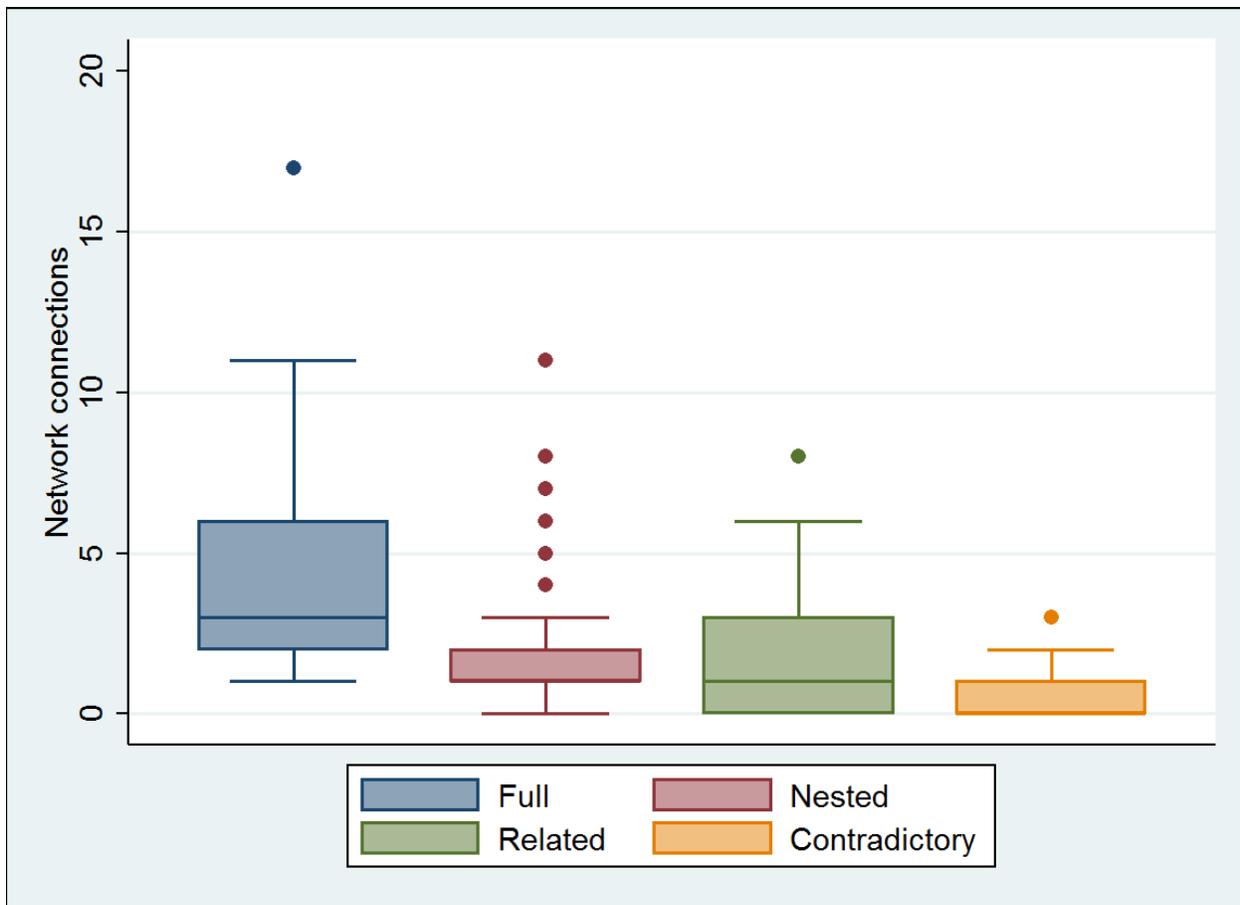
| Field | | Legend | | |
|--|---|-------------|---|---|
|  Collective action and the commons |  Interdisciplinary | Link |  Related | |
|  Political economy |  Conservation biology | | |  Contradictory |
|  Resilience |  Geography and land use change | | |  Nested |
|  Environmental & resource economics |  Political ecology | | | |

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Figure 5: Legend for theory-to-theory network diagrams

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Figure 6: Distributions of theory-to-theory connections by connection type

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Table 1: Summary of variable roles

| Type | Subtype | Description |
|-------------|----------------|--|
| Independent | Underlying | An independent variable that affects an outcome by affecting another, more proximate cause. Also referred to as a distal cause. |
| | Proximate | An independent variable that directly affects an outcome without the help of an intermediary variable. |
| | Moderating | An independent variable that affects an outcome by affecting the relationship between another independent variable and this outcome. This creates what is commonly referred to as an "interaction effect." |
| Outcome | Intermediate | An outcome that is affected by the independent variables in a theory, that in turn affects the final outcome (e.g. collective action of a commons user group) |
| | Final | The final outcome in a theory (e.g. the condition of an environmental commons as it is affected by levels of the intermediate outcome such as collective action) |

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Table 2: Types of theory-to-theory relationships.

| Relationship | Transitivity | Symmetry | Conditions and notes |
|---------------|--------------|------------|---|
| Nested | Fully | Asymmetric | 1) Theory A contains all of the variables that theory B contains, with either the same* values or the opposite* set of values. 2) These variables do not need to have precisely the same roles in the larger theory as they do in the nested one (e.g. a proximate cause may become an intermediate outcome depending on its place in the larger theory). |
| Related | Partially | Symmetric | 1A) Theory A and B have the same* value for the same independent variable and predict the same value for the same final outcome. 1B) OR theory A and theory B have opposite* values for the same independent variable and thus predict opposite values for the same final outcome, 2) AND the two theories do not share the common independent variable via a shared theory that is nested within each. |
| Contradictory | Not | Symmetric | 1) Theory A and B have the same* value for the same independent variable but predict a different value for the same final outcome. 2) Because of the principle of equifinality, theories that have different values for the same independent variable and the same value for the same final outcome are not considered to be contradictory. |

734 The table shows three types of relationships. These relationships vary in the extent to which the relationships involved are symmetric,
735 and in the extent to which they confer transitive relationships among theories. The only asymmetric relationship type is nested, in that A
736 containing B does not imply that B contains A. Otherwise the relationships are symmetric. The nested relationship is fully transitive: if A
737 contains B and B contains C, then theory A contains theory C and all of its variables. Moreover, because theory A contains all of the
738 variables (with the same, or when appropriate, opposite, values) as theory B, this also means that theory A inherits all of the other
739 relationships held by theory B. These inherited relationships are implicit in the variables that theory A contains and are not formally
740 coded for theory A. For example, if theory A contains theory B and theory B contains theory C and is contradictory with theory D, we did
741 not code that theory A had any formal relationship with theories C or D. Related relationships are partially transitive. If theory A is related
742 to theory B and theory B is related to theory C, whether or not theory A is then related to theory C depends on whether this relationship
743 also satisfies the conditions for the related association. Theories that are related to the same theory are more likely to be related to each
744 other in this way, but are not always related. Contradictory relationships are not transitive at all. However, if A and B are both
745 contradictory to C, then it is likely that they are related to each other.

746 *Two variables have the same value if the value each of is on the same side of the median value (e.g. both are “high” for an ordinal
747 variable). They have opposite values of their values are on the opposite side of the median value.
748

Table 3: Numerical summary of theories

| Attribute | Definition | Mean | Median | SD |
|----------------------|--|-------------|---------------|-----------|
| Variables | Number of variables used | 5.63 | 4.00 | 3.79 |
| Connections | Number of connections to other theories | 4.35 | 3.00 | 3.21 |
| Related | Number of "related" connections to other theories | 1.78 | 1.00 | 2.06 |
| Nested in | Number of theories a theory is nested in | 0.94 | 0.00 | 1.92 |
| Contains | Number of theories a theory contains | 0.94 | 1.00 | 1.56 |
| Contradictory | Number of "contradictory" connections to other theories | 0.70 | 0.00 | 0.99 |
| Steps | Number of steps (up to 4) a theory contains in its process | 2.97 | 3.00 | 0.78 |
| List | Maximum number of variables at one step a theory contains | 2.29 | 1.00 | 2.58 |
| Moderators | Number of moderator variables a theory contains | 0.95 | 0.00 | 1.56 |

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Table 4: Results by scientific field

| Field | Description and focus | Theories | Primary link type | Final outcome | Intermediate outcomes | Main Independent variable |
|-----------------------------------|---|----------|-------------------|-------------------------|--------------------------------------|---------------------------|
| Collective Action and the Commons | Explain the outcomes of emergent collective action among natural resource-users. | 21 | Nested | Commons condition trend | Collective action | Transaction costs |
| Political Economy | Explain how institutions and political contexts affect environmental outcomes across multiple scales | 14 | Related | Commons condition trend | Compliance and Social-ecological fit | Centralization |
| Interdisciplinary | Theories that were closely related to multiple disciplines | 7 | Nested | Commons condition trend | Compliance and Social-ecological fit | Governance knowledge use |
| Resilience | Explain conditions under which ecosystems retain their current structure and function and avoid a phase shift to an alternative stable state in the face of <i>unspecified</i> perturbation | 7 | Nested | Basin switch | Ecological resilience | Actor adaptive capacity |
| Environmental Economics | Focus on environmental policy instruments | 6 | Related | Commons condition trend | None | Policy instrument |
| Conservation Biology | Focus on conservation-based policies and governance regimes (e.g. protected areas) | 5 | Related | Commons condition trend | Compliance | Policy instrument |
| Geography and Land Use | Focus on spatial land use dynamics as affected by institutions and technology | 2 | Related | Commons condition trend | None | Technology role |
| Political Ecology | Focus on power dynamics and how these affect social and environmental outcomes | 1 | Related | Commons condition trend | Economic dependence | N/A |

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A scientific field is characterized by a suite of factors, namely: (1) a topic of study with a characteristic set of research questions and methods designed to answer these questions; (2) an identifiable community of scholars, usually organized around membership in one or more professional or academic societies, a set of conferences, and the set of journals they publish in.”

Table 5: Connections between theories by field

| | Commons | Con. Biology | Env. Economics | Geography | Interdisciplinary | Political ecology | Political economy | Resilience |
|--------------------------|----------------|---------------------|-----------------------|------------------|--------------------------|--------------------------|--------------------------|-------------------|
| Commons | 27 | | | | | | | |
| Con. Biology | 3 | 9 | | | | | | |
| Env. Economics | 0 | 0 | 1 | | | | | |
| Geography | 0 | 0 | 3 | 0 | | | | |
| Interdisciplinary | 8 | 3 | 4 | 0 | 6 | | | |
| Political ecology | 0 | 0 | 0 | 0 | 0 | 0 | | |
| Political economy | 13 | 10 | 3 | 1 | 12 | 4 | 12 | |
| Resilience | 5 | 0 | 1 | 1 | 3 | 0 | 4 | 4 |

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Table 6: Correlation matrix of theory attributes

| | Variables | Connections | Related | Nested in | Contains | Contradictory | Steps | List | Moderators |
|----------------------|------------------|--------------------|----------------|------------------|-----------------|----------------------|--------------|-------------|-------------------|
| Variables | 1.00 | | | | | | | | |
| Connections | 0.44 | 1.00 | | | | | | | |
| Related | 0.27 | 0.66 | 1.00 | | | | | | |
| Nested in | -0.26 | 0.28 | -0.32 | 1.00 | | | | | |
| Contains | 0.73 | 0.43 | 0.07 | -0.18 | 1.00 | | | | |
| Contradictory | 0.22 | 0.65 | 0.58 | -0.07 | 0.02 | 1.00 | | | |
| Steps | 0.36 | 0.11 | 0.07 | -0.16 | 0.29 | 0.09 | 1.00 | | |
| List | 0.87 | 0.48 | 0.21 | -0.19 | 0.85 | 0.14 | 0.30 | 1.00 | |
| Moderators | 0.31 | -0.02 | 0.06 | -0.19 | 0.04 | 0.10 | -0.40 | 0.00 | 1.00 |

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1. Collective action and the commons

Collective action and the commons, or the literature on common pool resources emerged during the 1980s and 1990s as a response to the view that embodied natural resource management until that time.

According to this view, sustainable use of natural resources could only be achieved through privatization or centralized management by an external authority. However, research on the topic of collective action and small-scale environmental management revealed that local groups of people, such as communities, can self-organize and sustainably manage their resources (1-3). These findings have had a significant impact on the theory and practice of natural resource management around the world.

Through the use of laboratory and field experiments, case studies, and large-n datasets, this body of literature identified over 30 variables potentially affecting sustainable resource management (4). Some of the more recent advances in the field of collective action and the commons try to assess how some of these variables interact to produce sustainable resource outcomes. This is evident in most of the theories in our database that are related to this field.

2. Political economy

Political economy is the study of the interrelationships between political and economic processes, and is thus pursued by a broad array of scholars from multiple disciplines and traditions. Within sustainability science, prominent political economic theories draw from both Marxist (see discussion under political ecology) and neoclassical roots (e.g., 18, 19, 20). Although these traditions work from different assumptions and have different concerns, they share a focus on understanding the impacts of institutions and governance on economic and ecological change, as well as understanding the ways that economic forces constrain institutional change.

Theories that we coded from this body of literature are in large part centered on the decentralization of natural resource management. Although decentralization as a concept has taken on different meanings in different contexts (21-24), we use it here to refer to the process of transferring authority over resources from central governments to lower level units. On the one hand, the literature on decentralization argues that these transfers in management authority are superior to centralized management approaches because greater local involvement in decision-making improves the fit between resources and users, aligning the incentives between the costs and benefits of resource management. On the other hand, it highlights the problems associated with decentralization by examining leakage and elite capture; all of which can result in the degradation of resources they are meant to protect.

3. Interdisciplinary

Several theories coded in our database could not be easily assigned to any one of the literatures described in this section. This is largely because the theories in this group constitute building blocks for many of the other theories in our database and cross disciplinary boundaries. For instance, enforcement is classified as an interdisciplinary theory because there is widespread agreement across most disciplines that enforcement plays an important role in generating desirable behavior and outcomes in social-ecological

systems. Given the close relatedness among the examined literatures in terms of their origins and theoretical underpinnings this situation is not entirely surprising.

4. Resilience

Resilience can be defined as the ability of a system to maintain its function, processes, and structure, and to persist in the presence of disturbance (6-8). This definition, according to Gunderson (7), implies the existence of multiple stable states of a given system. Once a resilience threshold has been reached, a system may lose its ability to maintain adequate structural processes and function and may switch to an alternate stable state. Despite significant global declines in environmental conditions, examples of systems that have undergone a basin switch remain relatively rare. This means that empirical evidence supporting resilience theory is derived from the few cases that have crossed resilience thresholds and switched to an alternative state (9) and from regimes that remain within their current configuration but that are seemingly better able to recover from disturbance events (10).

A more recent shift towards resilience theories underpinned by social-ecological systems science has led to a burgeoning of interdisciplinary research in this area and the positing of many principles for enhancing ecological or SES resilience (e.g., 11, 12). Some principles like the “Rule of Hand” which posits that at most five slow variables drive a particular system’s resilience are difficult to formalize because these slow variables are context dependent and specific to a particular system and set of perturbations (13). Regardless of these challenges, resilience theories have substantially contributed to natural resource governance by informing more adaptive, flexible forms of environmental management and by altering key management practices, such as fire management regimes in forests and savannahs (11) or the systematic design of protected area networks (14).

5. Environmental and resource economics

Environmental and resource economics is a distinct branch of economics that utilizes economic theory and methods to evaluate both the economy and the environment in which it is embedded. Evaluations are used to develop policy prescriptions for environmental problems such as pollution, waste, overharvesting, and to measure the costs and/or benefits that result from such activities. A common theme of environmental and resource economics is to incorporate costs associated with natural resource pollution or depletion into the model (i.e. externalities) and in that way correct price signals in the most efficient manner. It relies on a number of policy instruments, such as taxes, emission licenses, quotas, or technology control among others to achieve this goal (5).

The theories we coded include a number of different policy instruments including information provision, market-based instruments, payment for ecosystem services, and taxes. The effects of these policy instrument are typically based upon a presumption that external monitoring and sanctioning mechanisms exist to influence compliance with those policy instruments. Two of the coded theories from this literature (Environmental Kuznets Curve and Rebound Effect) do not include policy instruments. Instead, they explain the role of economic status and technological advances, respectively, on the use of the commons.

6. Conservation biology

Conservation biology is a mission-driven discipline, aiming to reduce or stop biodiversity loss caused by human activities (25). Key threats to biodiversity include direct mortality by people (e.g., hunting, fishing), habitat degradation, pollution, and climate change, among others (26, 27). Although conservation biology is deeply rooted in Western science and thought (28, 29), it also reflects and incorporates some of the traditional ideas and knowledge related to resource conservation and ecology in general (30). The main approach by which conservation biology seeks to promote biodiversity conservation is through protection of diverse habitats and maintenance of key ecological processes. Within this context, selection and protection of geographic areas that have high biodiversity value is of utmost importance.

The theories we coded from this body of literature rely on protected areas as a main policy instrument, although they apply this instrument within different governance models. One theory is that effectively designed networks of conservation areas will be effective at protecting biodiversity (comprehensive, adequate, representative principles for conservation area design). Similarly, another identified characteristics of effectiveness of marine protected areas (MPAs) (ecological effectiveness of MPAs). Earlier conservation views treated humans as an exogenous component of the natural system that should, to the highest extent possible, be removed from the conservation initiatives (i.e. fortress conservation, or centralized conservation). Based on largely negative outcomes of those early conservation efforts, more recent theories view humans and human-related activities as an important determinant of successful conservation outcomes (critique of fortress conservation). Similarly, another theory posits a relationship between positive cultural and livelihood impacts and positive conservation outcomes (local livelihood and protected areas). An important distinction among these latter three theories is in their depiction of humans' role in resource and biodiversity conservation.

7. Geography and land use change

Land use change literature investigates how human activities, such as forestry and agriculture, transform the landscape and in that way affect the natural exchange rate of greenhouse gas emissions between the atmosphere and land (31, 32). Such transformations and alterations of the landscape can potentially have substantial impact on regional and global climates (33).

The theories that we coded from this body of literature focus on the role of technology and economic status in reducing deforestation. In the agricultural context, technological improvements lead to higher productivity per unit of arable land and thus minimize conversion of forests to agricultural land. The effect of a rising economic status on the level of deforestation follows a Kuznets curve.

8. Political ecology

Political Ecology (PE) is an emerging supra-disciplinary approach to study human-environment interactions with a critical lens. PE analyzes the causes and outcomes (incl. the distribution of costs and benefits) of environmental changes in relation power relations in turn associated to the existing political economy (15-17). A central feature of political ecology is the *politicization* of environmental problems.

This means that environmental problems are understood as rooted in political clashes over alternative futures and clashes between alternative values and imaginaries. Therefore, there is a crucial distinction between what PE calls *political* explanations of environmental problems and *apolitical* ones. Apolitical explanations ‘naturalize’ certain conditions; these include explanations that mainly attribute environmental changes to geographical factors, population, poverty, technology and market failure. It is arguably not (mainly) an academic discipline but *an approach* to doing research and writing which shares some central propositions (theses).

Political ecology consists of five key “critical tools” or fields of study, which serve as pillars and which contribute important concepts and theories (17): Marxist political economy (also “green materialism”); peasant studies; feminist development studies; critical environmental history; and postcolonial studies. Drawing from these critical tools, PE has developed four main “theses” (or for the purpose of this paper, theories) which compose the broad research agenda of the field: (1) the systematic relation between processes of social-ecological marginalization and degradation; (2) the link between the unequal appropriation and metabolization of ecological resources for economic activities, socially-produced scarcities, and the emergence of environmental conflicts (also linked to the concept of environmental injustice); (3) the relation between environmental conservation initiatives and governmental or corporate control over territories and peoples; and (4) the role of social movements in achieving more just and sustainable socio-ecological conditions (17).

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Figure 1: Histograms of the number of variables used by theories and theories using variables.

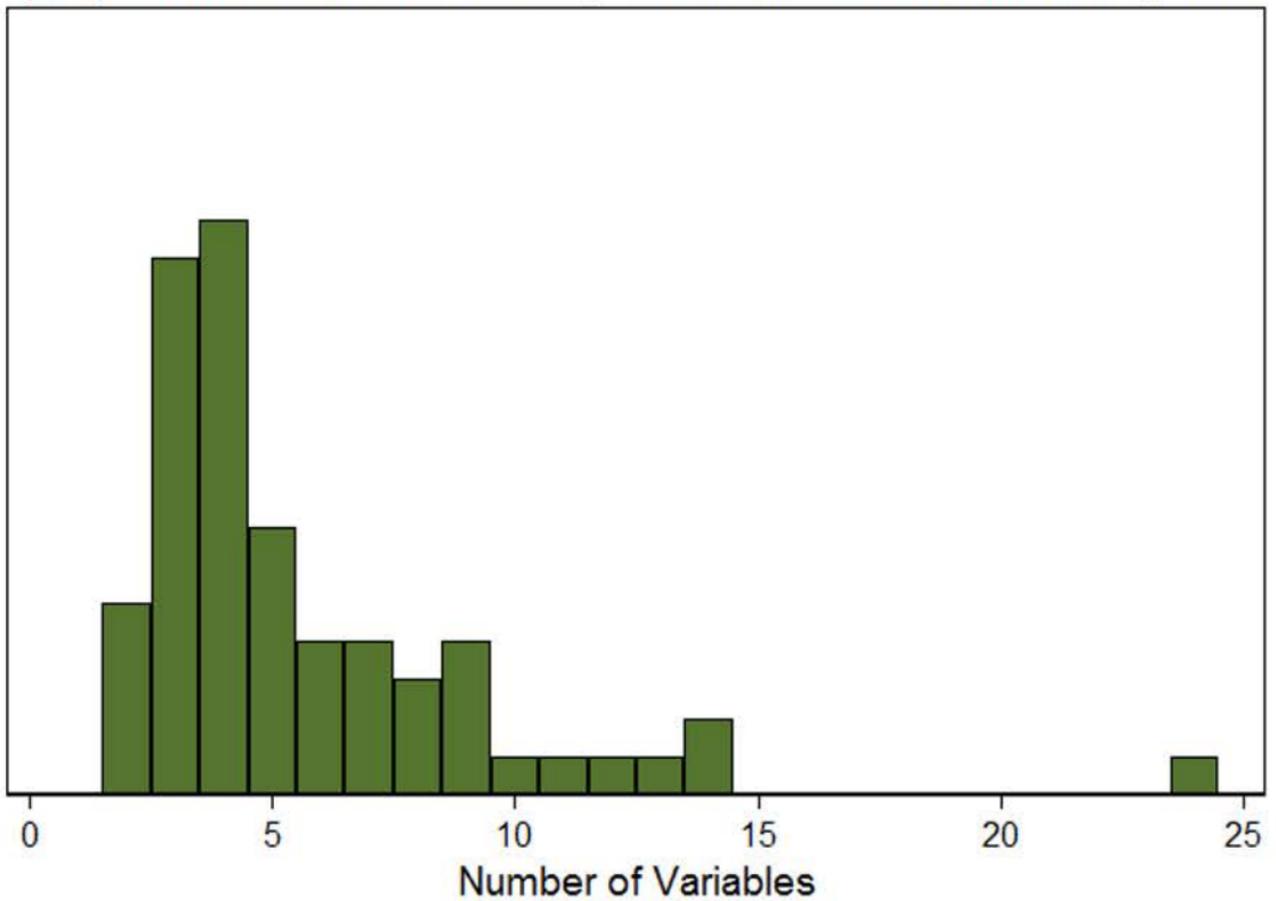
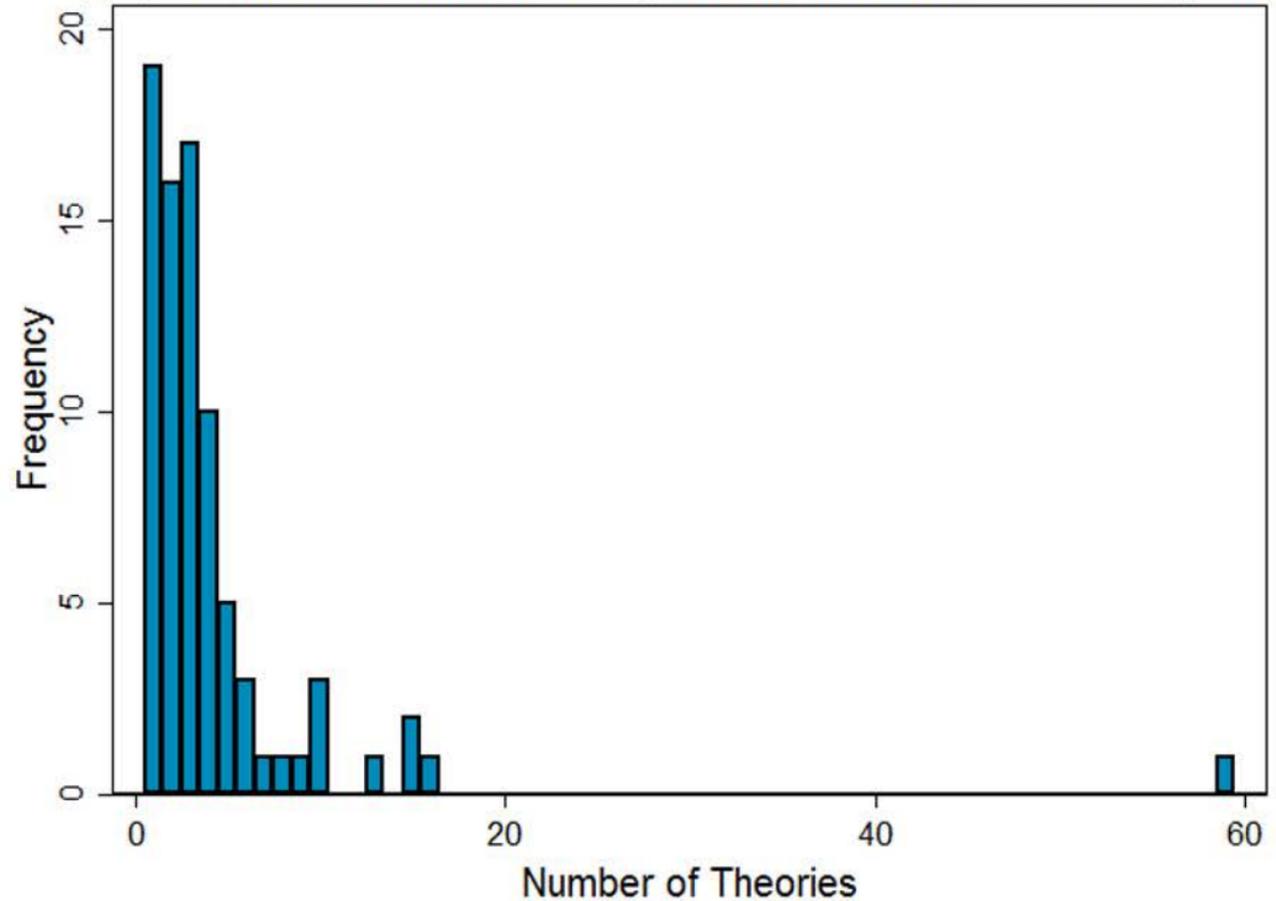
Figure 2: Theory-to-theory network of related and contradictory relationships

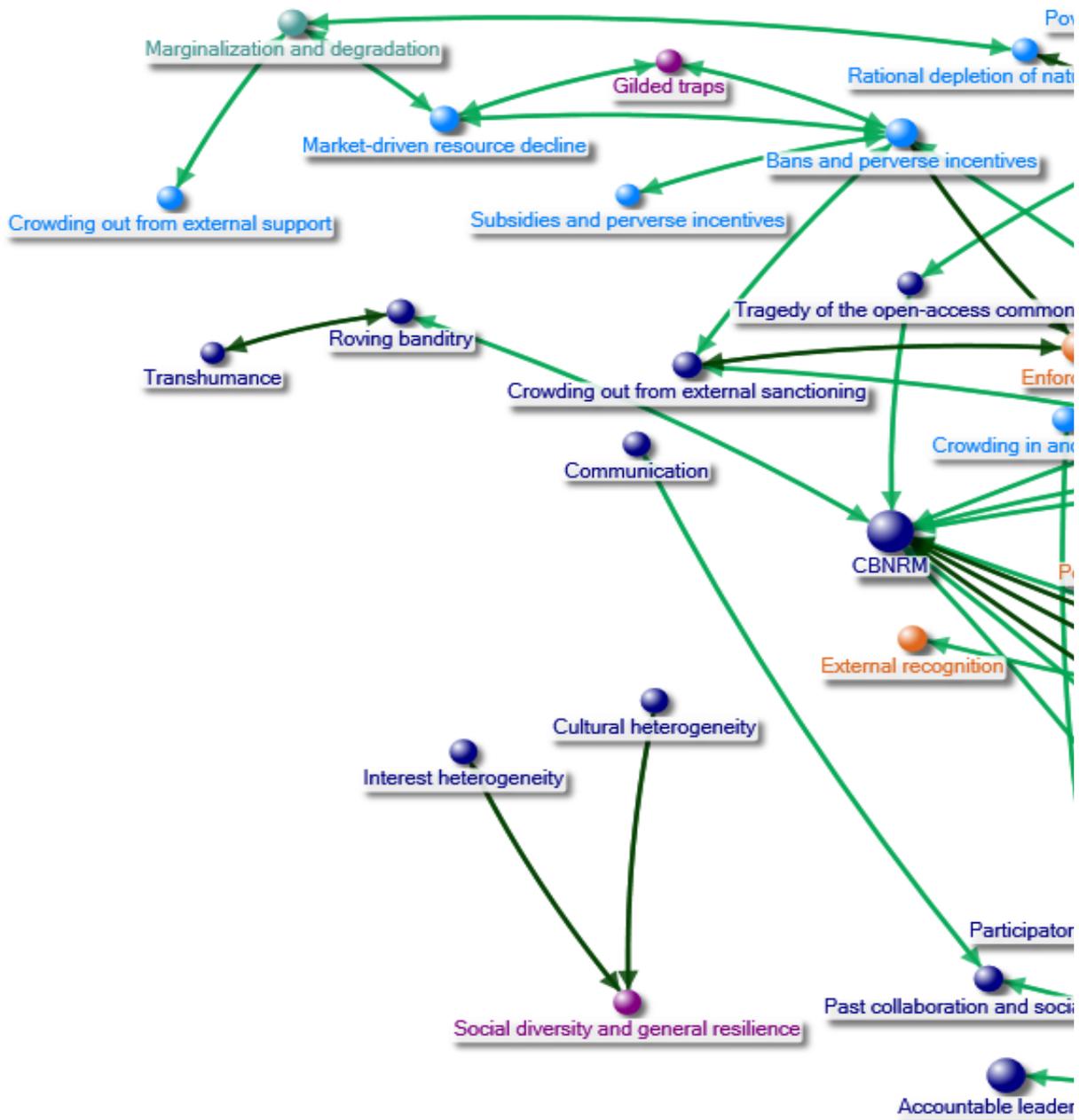
Figure 3: Theory-to-theory network with nested relationships

Figure 4: Theory-to-theory network with all relationships

Figure 5: Network legend

(1A) Distribution of variables by the number of theories using them (1B) Distribution of theories by the number of variables they use





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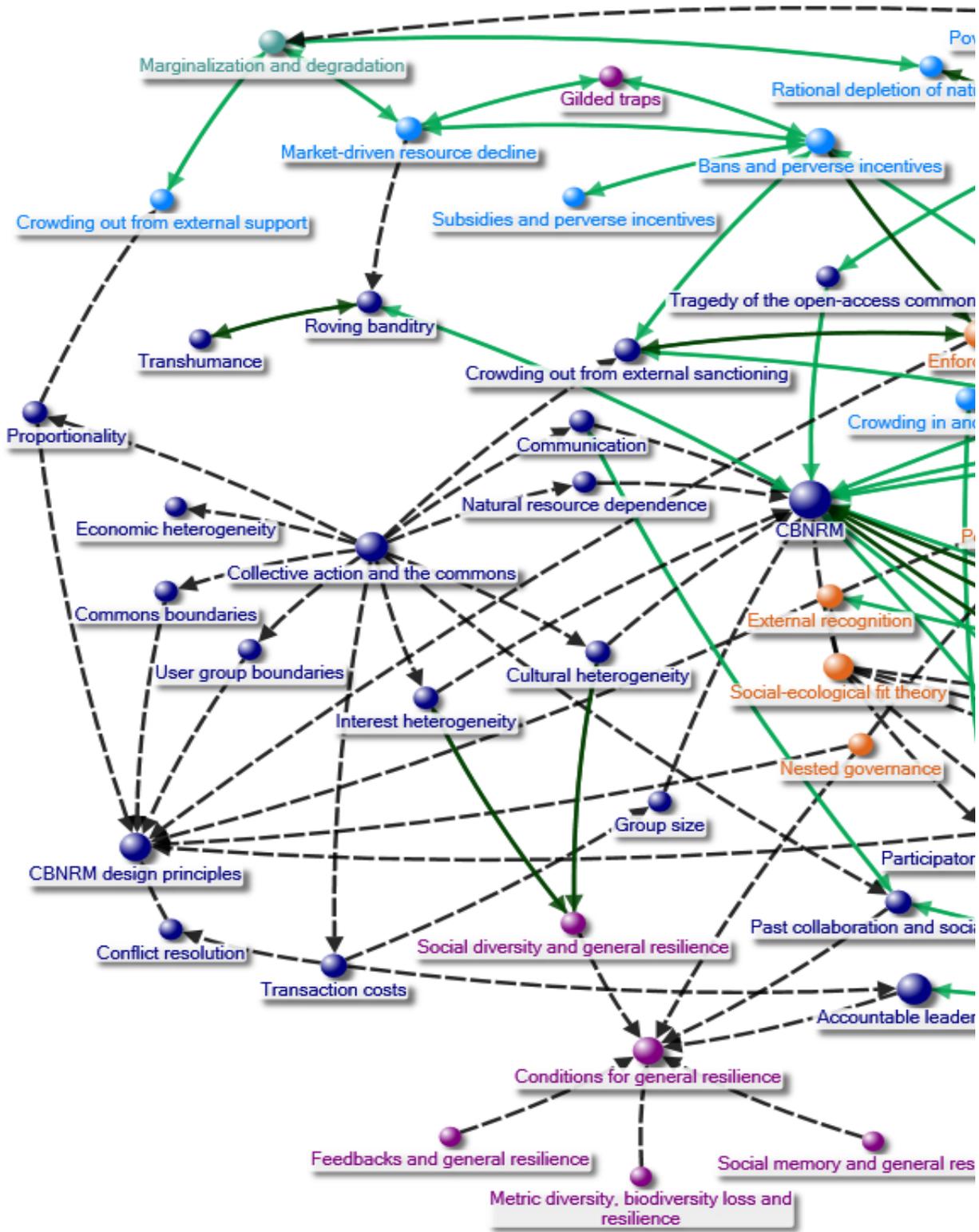
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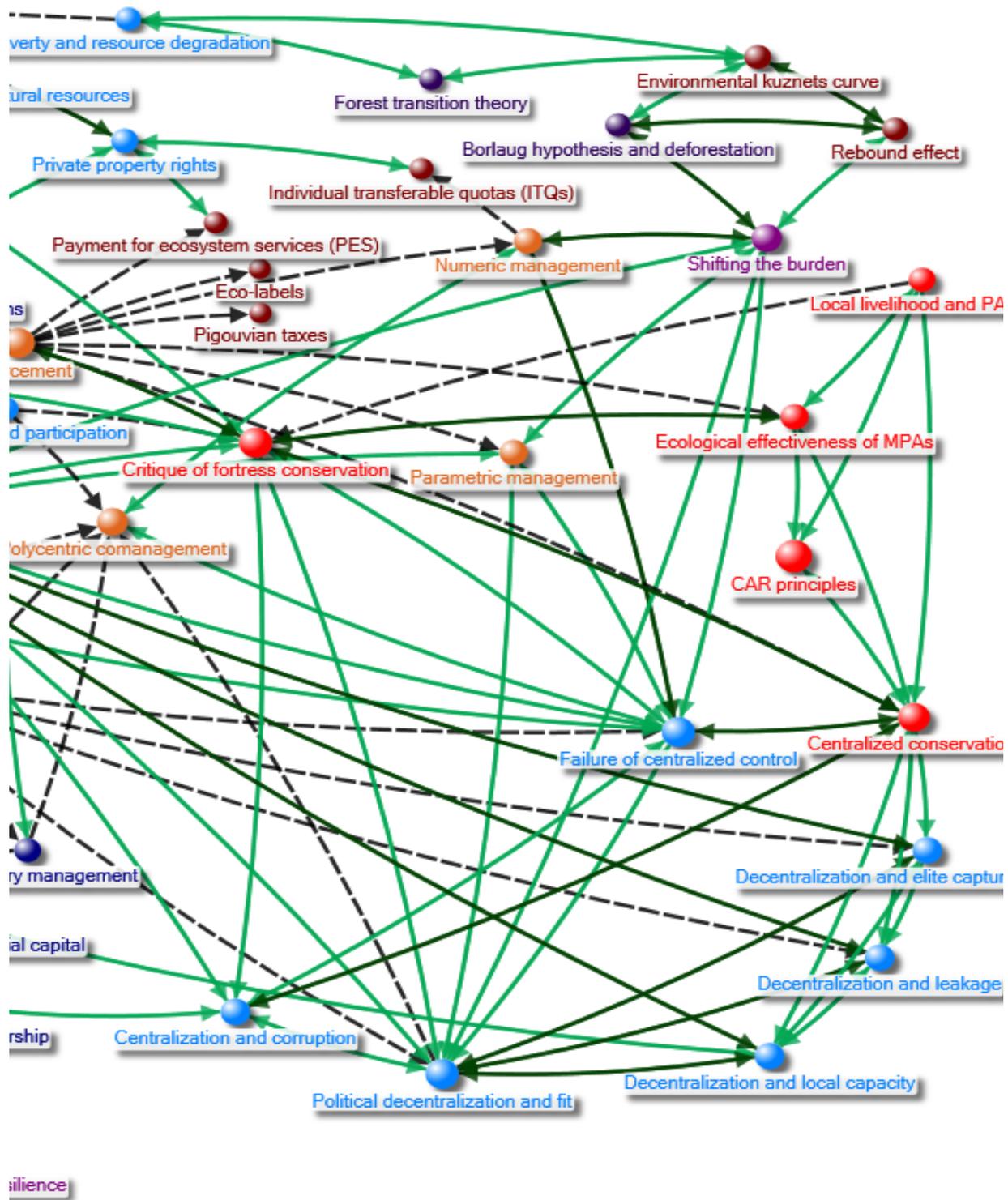
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| Field | | | | Legend | | |
|---|------------------------------------|---|-------------------------------|-------------|--|---|
|  | Collective action and the commons |  | Interdisciplinary | Link | | |
|  | Political economy |  | Conservation biology | | |  Related |
|  | Resilience |  | Geography and land use change | | |  Contradictory |
|  | Environmental & resource economics |  | Political ecology | | |  Nested |

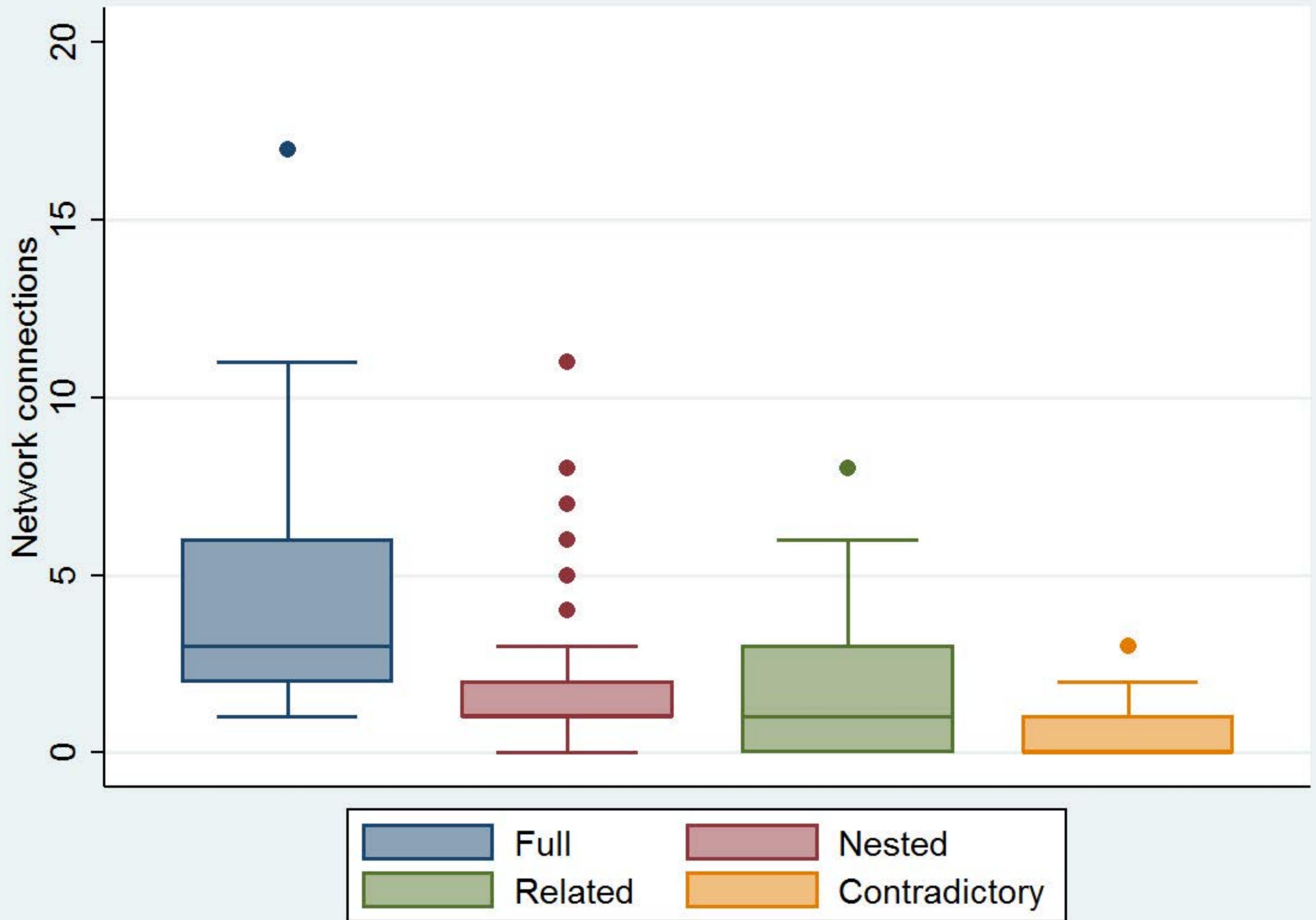


Table 1: Summary of variable roles

| Type | Subtype | Description |
|-------------|----------------|--|
| Independent | Underlying | An independent variable that affects an outcome by affecting another, more proximate cause. Also referred to as a distal cause. |
| | Proximate | An independent variable that directly affects an outcome without the help of an intermediary variable. |
| | Moderating | An independent variable that affects an outcome by affecting the relationship between another independent variable and this outcome. This creates what is commonly referred to as an “interaction effect.” |
| Outcome | Intermediate | An outcome that is affected by the independent variables in a theory, that in turn affects the final outcome (e.g. collective action of a commons user group) |
| | Final | The final outcome in a theory (e.g. the condition of an environmental commons as it is affected by levels of the intermediate outcome such as collective action) |

Table S2: Types of theory-to-theory relationships

| Relationship | Transitivity | Symmetry | Conditions and notes |
|---------------|--------------|------------|---|
| Nested | Fully | Asymmetric | 1) Theory A contains all of the variables that theory B contains, with either the same* values or the opposite* set of values. 2) These variables do not need to have precisely the same roles in the larger theory as they do in the nested one (e.g. a proximate cause may become an intermediate outcome depending on its place in the larger theory). |
| Related | Partially | Symmetric | 1A) Theory A and B have the same* value for the same independent variable and predict the same value for the same final outcome, 1B) OR theory A and theory B have opposite* values for the same independent variable and thus predict opposite values for the same final outcome, 2) AND the two theories do not share the common independent variable via a shared theory that is nested within each. |
| Contradictory | Not | Symmetric | 1) Theory A and B have the same value for the same independent variable but predict a different value for the same final outcome. 2) Because of the principle of equifinality, theories that have different values for the same independent variable and the same value for the same final outcome are not considered to be contradictory. |

The table shows three types of relationships. These relationships vary in the extent to which the relationships involved are symmetric, and in the extent to which they confer transitive relationships among theories. The only asymmetric relationship type is nested, in that A containing B does not imply that B contains A. Otherwise the relationships are symmetric.

The nested relationship is fully transitive: if A contains B and B contains C, then theory A contains theory C and all of its variables. Moreover, because theory A contains all of the variables (with the same, or when appropriate, opposite, values) as theory B, this also means that theory A inherits all of the other relationships held by theory B. These inherited relationships are implicit in the variables that theory A contains and are not formally coded for theory A. For example, if theory A contains theory B and theory B contains theory C and is contradictory with theory D, we did not code that theory A had any formal relationship with theories C or D.

Related relationships are partially transitive. If theory A is related to theory B and theory B is related to theory C, whether or not theory A is then related to theory C depends on whether this relationship also satisfies the conditions for the related association. Theories that are related to the same theory are more likely to be related to each other in this way, but are not always related. Contradictory relationships are not transitive at all. However, if A and B are both contradictory to C, then it is likely that they are related to each other.

*Two variables have the same value if the value each of is on the same side of the median value (e.g. both are “high” for an ordinal variable). They have opposite values if their values are on the opposite side of the median value.

Table S3: Summary of variables

| Variable | Component type | Type | Theories | Question | Range |
|--------------------------------|-----------------------|-------------|----------|---|---|
| Commons Condition Trend | Environmental commons | Ordinal | 59 | Based on your answers to the Beginning Condition and End Condition variables, would you say that the condition of this commons has improved, remained the same, or worsened during this snapshot? | 1 Worsened; 2 Remained the Same; 3 Improved |
| Collective Action | Actor group | Ordinal | 16 | What is the current level of collective action within the members of this actor group with respect to the use or management of this commons? | 1 Low; 2 Medium; 3 High |
| Compliance | Actor group | Ordinal | 15 | Do members of this actor group follow the rules of this governance system with respect to the emission or appropriation of this commons? | 1 No; 2 Somewhat; 3 Yes |
| Transaction Costs | Governance system | Ordinal | 15 | How high (or low) are the transaction costs of monitoring and enforcing the rules that this governance system involves in managing this commons? | 1 Low; 2 Medium; 3 High |
| Policy Instrument | Governance system | Categorical | 13 | Does this formal governance system apply any of the following policy instruments to this commons? | Proportional outcome-based performance standard; Absolute outcome-based performance standard; Technological prohibition; Technological mandate; Temporal standard; Ban; Price ceiling ; Price floor ; Tax ; Subsidy; PES scheme; Joint tax-subsidy; Market-based instrument; Information provision; Insurance provision; Protected area |
| Centralization | Governance system | Ordinal | 10 | Is this governance system highly centralized or highly decentralized? | 1 Highly decentralized; 2 Somewhat decentralized; 3 Somewhat centralized; 4 Highly centralized |

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| External sanctions | Actor group | Binary | 10 | Are sanctions applied by other actor groups to the members of this group for violations of rules regarding extraction or emission? And if so, are these sanctions graduated (increasing with severity and repetition of offenses)? | Yes; No |
| Social-Ecological Fit | Governance system | Ordinal | 10 | To what extent (low, medium, or high) do the institutional arrangements of this governance system fit well with the ecological or physical features of the commons on which they are implemented? | 1 Low; 2 Medium; 3 High |
| External Monitoring | Actor group | Binary | 9 | Do external actors/organizations monitor the activities of this commons using actor group with respect to the use of this commons? | Yes; No |
| Governance Knowledge Use | Governance system | Categorical | 8 | What type(s) of knowledge does this governance system employ in its management of the commons it governs? | Scientific knowledge; Local/traditional knowledge |
| Actor Adaptive Capacity | Actor group | Ordinal | 7 | How would you rate the adaptive capacity of this actor group with respect to large changes in the availability or concentration of the commons they rely on in this snapshot? | 1 Low; 2 Medium; 3 High |
| Ecological Resilience | Environmental commons | Ordinal | 6 | Given the current state of the system, how ecologically resilient is this commons to the threats that it can be expected to face? | 1 Poorly resilient; 2 Moderately resilient; 3 Highly resilient |
| Property Regime | Actor group | Categorical | 6 | What property regime does this actor group apply to this commons? | Private property; Common property; Public property; Corporate property; Open-access |
| Actor Traditional Knowledge | Actor group | Ordinal | 5 | What is the level of traditional or local knowledge this actor group has regarding the condition of this environmental commons? | 1 Low; 2 Medium; 3 High |
| Basin Switch | Environmental commons | Categorical | 5 | Does this natural resource show evidence of switching stable states during this snapshot? If not, is the current stable state considered to be in a desirable / undesirable state? If yes, is the new stable state considered to be desirable / undesirable? | Yes desirable; Yes undesirable; No desirable; No undesirable; Unclear - system may be transitioning; N/A |
| Commons Feedback Visibility Use | Actor group | Ordinal | 5 | How visible are the effects of commons use to this actor group? | 1 Low; 2 Medium; 3 High |

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|-------------------------------------|-----------------------|-------------|---|--|--|
| Metric Diversity | Governance system | Ordinal | 5 | Does this governance system focus exclusively on one target outcome, or on a range of metrics that are interrelated? | 1 Low: One metric for success; 2 Medium: Few metrics for success; 3 High: Many metrics for success |
| Participation in Rule Making | Actor group | Ordinal | 5 | How high is the level of participation of this actor group in the process that determines how this environmental commons is governed? | 1 Low; 2 Medium; 3 High |
| Actor Group Trust | Actor group | Ordinal | 4 | How high is the level of trust between members of this group? | 1 Low; 2 Medium; 3 High |
| Actor Scientific Knowledge | Actor group | Ordinal | 4 | What is the level of scientific knowledge this actor group has regarding the condition of this environmental commons? | 1 Low; 2 Medium; 3 High |
| Cultural Heterogeneity | Actor group | Ordinal | 4 | How high is the level of variation in the cultural identity of the group members? | 1 Low; 2 Medium; 3 High |
| Economic Dependence | Actor group | Ordinal | 4 | How dependent are the members of the group on this commons for their economic well-being? | 1 Not dependent or Slightly dependent; 2 Moderately dependent; 3 Very dependent |
| Economic Status | Actor group | Ordinal | 4 | What is the average economic status of the group? (Does not refer to variation within the group – see ActorEconomicHeterogeneity). | 1 Low; 2 Medium; 3 High |
| External Recognition | Governance system | Ordinal | 4 | Within this governance system, do larger governmental jurisdictions (i.e. International agreements, Nation states) recognize the autonomy of lower-level jurisdictions (States, regions, communities), and their right to make decisions regarding this commons? | 1 Low - no recognition; 2 Moderate - some recognition; 3 High - complete recognition |
| Interest Heterogeneity | Actor group | Ordinal | 4 | How much do the interests of the members of this group diverge from each other? | 1 Low; 2 Medium; 3 High |
| Multiple Levels | Governance system | Categorical | 4 | Does this governance system contain multiple levels, with each level having a set of actors who conduct tasks with respect to the management of this commons? If so, is there active coordination across these levels, or not? | Single-level governance; Coordination among multiple levels; Multiple levels but no coordination |
| Productivity | Environmental commons | Ordinal | 4 | How productive is the commons? For renewable commons this is often positively associated with levels of rainfall and solar radiation. (e.g., Amazon rainforests have very high | 1 Poorly productive; 2 Moderately Productive; 3 Very productive |

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| | | | | productivity, while deserts have low productivity). | |
| Resource Market Value | Environmental commons | Ordinal | 4 | If it is traded in a market, how high is the market value of this resource? | 1 Low; 2 Medium; 3 High |
| Technology Role | Actor group | Categorical | 4 | How has technology affected the relationship between this actor group and the commons it uses in this interaction? | Increased commons conservation; Increased commons use; Increased productivity |
| Biodiversity Trend | Environmental commons | Ordinal | 3 | What is the trend in biodiversity of this natural resource system during the time frame for which the governance system has been in place? | 1 Worsened; 2 Mixed effects or remained the same; 3 improved |
| Commons Alternatives | Actor group | Ordinal | 3 | In addition to this particular commons, does this actor group have access to other commons of the same type (e.g. if the commons in use is a forest, other forests) that would serve the same function? | 1 Easily access other commons; 2 Can access other commons with some difficulty; 3 Cannot access other commons |
| Commons Feedback Speed Use | Actor group | Ordinal | 3 | How quickly does this actor group experience the effects of its own use of this environmental commons? | 1 Low; 2 Medium; 3 High |
| Commons User Mobility | Actor group | Ordinal | 3 | Does this commons user group have a large amount of mobility to move within this system and between this system and other systems? | 1 Low; 2 Medium; 3 High |
| Conflict Resolution | Actor group | Binary | 3 | Are mechanisms in place to address conflicts that arise over the use of this commons by this actor group? | Yes; No |
| External Support | Governance system | Ordinal | 3 | Within this governance system, do larger governmental and/or non-governmental organizations actively support (e.g. through the supply of physical or financial resources, information) lower level jurisdictions (States, Regions, Cities)? | 1 No support; 2 Some support; 3 Extensive support |
| Institutional Diversity | Governance system | Ordinal | 3 | How diverse are the institutions that are implemented by this governance system on this commons? Do these institutions vary systematically with natural variations in properties of this commons? | 1 Low; 2 Medium; 3 High |
| Leadership Accountability | Actor group | Ordinal | 3 | How accountable are the leaders to the other members of the | 1 Low; 2 Medium; 3 High |

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| | | | | group? | |
| Market Scale | Environmental commons | Categorical | 3 | If there are markets for goods produced by or associated with this resource, what is the largest scale at which they operate? | local; sub-national; national; international; global; other |
| Markets | Environmental commons | Binary | 3 | Are there well-articulated markets for goods produced by or associated with this commons that affect this particular commons? | Yes; No |
| PA IUCN Strict Zones | Governance system | Interval | 3 | What percentage of the area of this PA is covered by no take zones (IUCN Ia, Ib, and II)? | Real numbers between 0 and 100 |
| Past Collaboration | Actor group | Ordinal | 3 | How high is the level of past collaboration within this group? | 1 Low; 2 Medium; 3 High |
| Property Security | Actor group | Ordinal | 3 | How secure are the rights that this commons user group has with respect to this environmental commons? | 1 Low; 2 Medium; 3 High |
| Proportionality (of costs and benefits) | Actor group | Binary | 3 | Is there general proportionality between the amount of costs group members incur and the amount of benefits received? | Yes; No |
| Self-Monitoring | Actor group | Binary | 3 | Does this actor group monitor its own activities with respect to the use of this commons? | Yes; No |
| Self-Sanctions | Actor group | Categorical | 3 | Are sanctions applied by and to the members of this group for violations of rules regarding extraction or emission? And if so, are these sanctions graduated (increasing with severity and repetition of offenses)? | Graduated sanctions; Non-graduated sanctions; No sanctions |
| User-Commons Proximity | Actor group | Binary | 3 | Does this actor group reside within or adjacent to the primary resource in this interaction? | Yes; No |
| Actor Group Boundary Clarity | Actor group | Categorical | 2 | Are there clear rules that are followed about who and who isn't a member of this group? | No boundaries; Unclear boundaries; Clear boundaries |
| Actor Group Size | Actor group | Interval | 2 | For this variable either enter the number of actors (e.g. 30), or if the number of actors is very large and essentially uncountable, enter "Many." | Positive whole numbers or "Many" |
| Actor Vulnerability | Actor group | Ordinal | 2 | How would you rate the vulnerability of this actor group to changes with respect to the commons that they depend on? | 1 Low; 2 Medium; 3 High |
| Commons Boundaries | Environmental commons | Ordinal | 2 | Are the boundaries that define the spatial extent of this commons clearly defined and highly visible? | 1 Very unclear boundaries; 2 Somewhat unclear boundaries; 3 Clear boundaries |

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|------------------------------------|-----------------------|-------------|---|--|--|
| Commons Political Power | Actor group | Ordinal | 2 | How much power does this actor group have in the process that determines the governance of this commons? | 1 Low; 2 Medium; 3 High |
| Cultural Services Condition | Environmental commons | Ordinal | 2 | What is the general trend in the condition of cultural services (e.g. spiritual, aesthetic, recreation, education) derived from this commons during the time frame of this snapshot? | 1 Worsened; 2 Mixed effects or remained the same; 3 Improving |
| Environmental Monitoring | Actor group | Ordinal | 2 | How much environmental monitoring of this commons does this actor group engage in? | 1 Low; 2 Moderate; 3 High |
| Leadership | Actor group | Categorical | 2 | What type of leadership does this group have, if any? | No leader; Formal leader; Informal leader |
| Leadership Authority | Actor group | Ordinal | 2 | How much authority does the leader of this group hold? | 1 Low; 2 Medium; 3 High |
| Leakage | Environmental commons | Categorical | 2 | Has the governance of this commons led to the leakage of costs or benefits onto other systems? In terms of costs, has the governance of this commons increased extraction or pollution pressures on other similar resources? | Yes, leakage of benefits; Yes, leakage of costs; No leakage |
| Livelihood Alternatives | Actor group | Ordinal | 2 | Other than this commons, does this actor group have access to alternative sources of economic livelihood or cultural well being? | 1 Easily access other alternatives; 2 Can access other alternatives with some difficulty; 3 Cannot access alternatives |
| Outsider Exclusion | Actor group | Ordinal | 2 | To what extent are members of this commons user group able to exclude non-members (outsiders) from using this commons? | 1 No exclusion; 2 Some exclusion; 3 Total exclusion |
| Over-Capitalization | Actor group | Binary | 2 | Is overcapitalization (e.g. excessive fishing capacity that overwhelms renewal rates of fish or another natural resource) a problem for this user group? | Yes; No |
| Personal Communication | Actor group | Ordinal | 2 | How frequently do the actors within this group, or individual representatives, communicate in person? | 1 Never; 2 Less than once every 2 years; 3 Once every 2 years; 4 Once a year; 5 More than once a year |
| Perverse Incentives | Actor group | Binary | 2 | Do higher levels of governance in this system provide this actor group incentives to over-appropriate or over-emit? For example, through subsidizing fishing fleets or agriculture crops? | Yes; No |

| | | | | | |
|---|-----------------------|-------------|---|--|---|
| User Group Well-Being Change | Actor group | Ordinal | 2 | How has the well-being of this commons user group changed during the time period identified in this interaction? | 1 Worsened; 2 Remained the Same; 3 Improved |
| Actor Debt | Actor group | Ordinal | 1 | Has this commons user group had to leverage their assets and take on debt in order to establish and continue their use of this commons? | 1 Little to no debt; 2 A moderate level of debt; 3 A high level of debt |
| Black Markets | Environmental commons | Binary | 1 | Are there well-articulated illegal markets for goods produced by or associated with this resource that affect this particular commons? | Yes; No |
| Causal Level | Governance system | Ordinal | 1 | Would you say that the governance system associated with this commons is addressing the more proximate causes, or addressing the more ultimate, underlying causes of the environmental and social problems that can be associated with this commons? | 1 Proximate; 2 Between proximate and ultimate; 3 Ultimate |
| Commons Boundary Negotiability | Actor group | Ordinal | 1 | How negotiable is access by non-members of this actor group to this environmental commons? | 1 Rigid ; 2 Moderate; 3 Negotiable |
| Commons Scarcity | Environmental commons | Ordinal | 1 | How scarce is this resource? | Not scarce -- abundant; Moderately scarce; Highly scarce |
| Economic Heterogeneity | Actor group | Ordinal | 1 | How heterogeneous are the members of this actor group in economic terms (wealth, income)? | 1 Low; 2 Medium; 3 High |
| External Ecological Connectivity | Environmental commons | Ordinal | 1 | How high is the level of ecological connectivity of this natural resource system? | 1 Low; 2 Medium; 3 High |
| Governance System Age | Governance system | Interval | 1 | How old is this governance system at the beginning of this interaction? | Positive real numbers |
| Governance System Spatial Extent | Governance system | Interval | 1 | What is the approximate spatial extent of this governance system (put in terms of square kilometers)? | Positive real numbers |
| Horizontal Coordination | Governance system | Categorical | 1 | What type of coordination do the members of this actor group engage in with members of other actor groups that are also involved in the use and/or management of the resource? | No coordination; Informal; Formal; Both formal and informal |
| Inter Annual Predictability | Environmental commons | Ordinal | 1 | How predictable is the availability or prevalence of this commons from year to year? | 1 Low; 2 Moderate; 3 High |
| Intra Annual Predictability | Environmental commons | Ordinal | 1 | How predictable is the availability or prevalence of this commons within years? | 1 Low; 2 Moderate; 3 High |

| | | | | | |
|---|-----------------------|-------------|---|---|---|
| Internal Ecological Connectivity | Environmental Commons | Ordinal | 1 | How high is the level of ecological connectivity between this natural resource system and its external environment? | 1 Low; 2 Medium; 3 High |
| MPA Internal Natural Boundaries | Governance system | Ordinal | 1 | Does the marine protected area protect an ecologically coherent area (i.e., limited or protected by deep water or sand) within no-take zones? | 1 Low; 2 Medium; 3 High |
| PA CAR Principles | Governance system | Ordinal | 1 | Were the ecological principles of Comprehensive, Adequate, Representative considered in the design of this MPA? | 1 No; 2 Partially; 3 Yes |
| Participation in PA Siting | Actor group | Ordinal | 1 | How high was the level of participation of this actor group or their representatives in siting of the marine protected area? | 1 Low; 2 Medium; 3 High |
| Rights Granting Process | Actor group | Categorical | 1 | Which of the following characterizes the process by which property rights to extract or emit this resource were granted to this actor group? | Based on historic extraction/emission rates; Based on current extraction/emission rates; Based on amount of land owned; Based on rights in another resource; Based on need; Based on political dynamics (power); Based on ethnic group; Based on gender ; Based on nationality; Based on sale or auction; Based on random assignment or lottery |
| Roving Bandit | Actor group | Binary | 1 | Would you say that this actor acts like a "roving bandit" (moving from place to place and degrading resources along the way)? | Yes; No |
| Scale Match | Governance system | Binary | 1 | Does the scale of this governance system match the scale of the commons that it is governing? | Yes; No |

Table 2: Types of theory-to-theory relationships

| Relationship | Transitivity | Symmetry | Conditions and notes |
|---------------|--------------|------------|---|
| Nested | Fully | Asymmetric | 1) Theory A contains all of the variables that theory B contains, with either the same* values or the opposite* set of values. 2) These variables do not need to have precisely the same roles in the larger theory as they do in the nested one (e.g. a proximate cause may become an intermediate outcome depending on its place in the larger theory). |
| Related | Partially | Symmetric | 1A) Theory A and B have the same* value for the same independent variable and predict the same value for the same final outcome, 1B) OR theory A and theory B have opposite* values for the same independent variable and thus predict opposite values for the same final outcome, 2) AND the two theories do not share the common independent variable via a shared theory that is nested within each. |
| Contradictory | Not | Symmetric | 1) Theory A and B have the same value for the same independent variable but predict a different value for the same final outcome. 2) Because of the principle of equifinality, theories that have different values for the same independent variable and the same value for the same final outcome are not considered to be contradictory. |

The table shows three types of relationships. These relationships vary in the extent to which the relationships involved are symmetric, and in the extent to which they confer transitive relationships among theories. The only asymmetric relationship type is nested, in that A containing B does not imply that B contains A. Otherwise the relationships are symmetric.

The nested relationship is fully transitive: if A contains B and B contains C, then theory A contains theory C and all of its variables. Moreover, because theory A contains all of the variables (with the same, or when appropriate, opposite, values) as theory B, this also means that theory A inherits all of the other relationships held by theory B. These inherited relationships are implicit in the variables that theory A contains and are not formally coded for theory A. For example, if theory A contains theory B and theory B contains theory C and is contradictory with theory D, we did not code that theory A had any formal relationship with theories C or D.

Related relationships are partially transitive. If theory A is related to theory B and theory B is related to theory C, whether or not theory A is then related to theory C depends on whether this relationship also satisfies the conditions for the related association. Theories that are related to the same theory are more likely to be related to each other in this way, but are not always related. Contradictory relationships are not transitive at all. However, if A and B are both contradictory to C, then it is likely that they are related to each other.

*Two variables have the same value if the value each of is on the same side of the median value (e.g. both are “high” for an ordinal variable). They have opposite values if their values are on the opposite side of the median value.

Table 3: Numerical summary of theories

| Attribute | Definition | Mean | Median | SD |
|----------------------|--|-------------|---------------|-----------|
| Variables | Number of variables used | 5.63 | 4.00 | 3.79 |
| Connections | Number of connections to other theories | 4.35 | 3.00 | 3.21 |
| Related | Number of "related" connections to other theories | 1.78 | 1.00 | 2.06 |
| Nested in | Number of theories a theory is nested in | 0.94 | 0.00 | 1.92 |
| Contains | Number of theories a theory contains | 0.94 | 1.00 | 1.56 |
| Contradictory | Number of "contradictory" connections to other theories | 0.70 | 0.00 | 0.99 |
| Steps | Number of steps (up to 4) a theory contains in its process | 2.97 | 3.00 | 0.78 |
| List | Maximum number of variables at one step a theory contains | 2.29 | 1.00 | 2.58 |
| Moderators | Number of moderator variables a theory contains | 0.95 | 0.00 | 1.56 |

Table 4: Results by scientific field

| Field | Description and focus | Theories | Primary link type | Final outcome | Intermediate outcomes | Main Independent variable |
|-----------------------------------|---|----------|-------------------|-------------------------|--------------------------------------|---------------------------|
| Collective Action and the Commons | Explain the outcomes of emergent collective action among natural resource-users. | 21 | Nested | Commons condition trend | Collective action | Transaction costs |
| Political Economy | Explain how institutions and political contexts affect environmental outcomes across multiple scales | 14 | Related | Commons condition trend | Compliance and Social-ecological fit | Centralization |
| Interdisciplinary | Theories that were closely related to multiple disciplines | 7 | Nested | Commons condition trend | Compliance and Social-ecological fit | Governance knowledge use |
| Resilience | Explain conditions under which ecosystems retain their current structure and function and avoid a phase shift to an alternative stable state in the face of <i>unspecified</i> perturbation | 7 | Nested | Basin Switch | Ecological resilience | Actor adaptive capacity |
| Environmental Economics | Focus on environmental policy instruments | 6 | Related | Commons condition trend | None | Policy instrument |
| Conservation Biology | Focus on conservation-based policies and governance regimes (e.g. protected areas) | 5 | Related | Commons condition trend | Compliance | Policy instrument |
| Geography and Land Use | Focus on spatial land use dynamics as affected by institutions and technology | 2 | Related | Commons condition trend | None | Technology role |
| Political Ecology | Focus on power dynamics and how these affect social and environmental outcomes | 1 | Related | Commons condition trend | Economic dependence | N/A |

A scientific field is characterized by a suite of factors, namely: (1) a topic of study with a characteristic set of research questions and methods designed to answer these questions; (2) an identifiable community of scholars, usually organized around membership in one or more professional or academic societies, a set of conferences, and the set of journals they publish in.

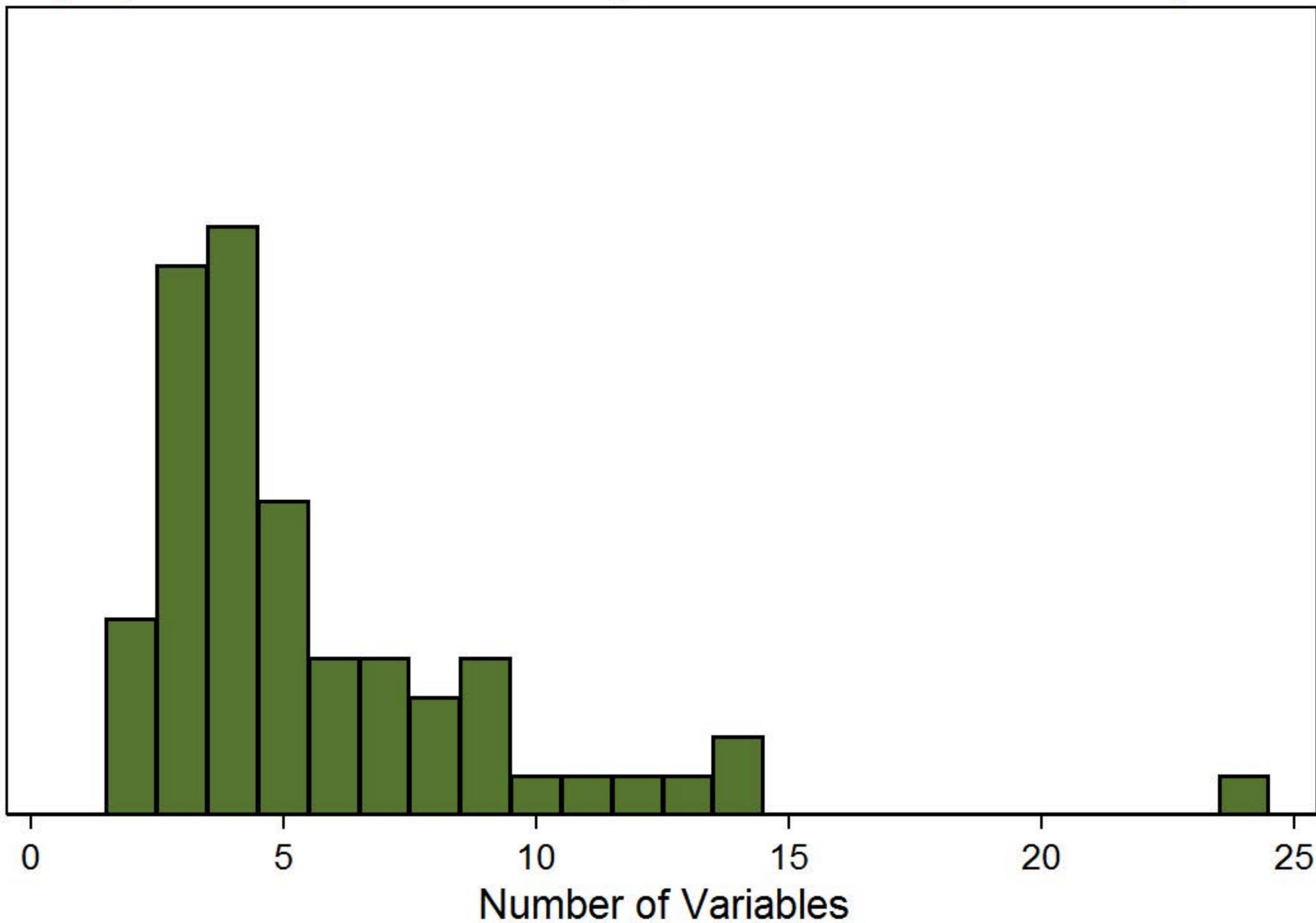
Table 4: Connections between theories by field

| | Commons | Con. Biology | Env. Economics | Geography | Interdisciplinary | Political ecology | Political economy | Resilience |
|--------------------------|----------------|---------------------|-----------------------|------------------|--------------------------|--------------------------|--------------------------|-------------------|
| Commons | 27 | | | | | | | |
| Con. Biology | 3 | 9 | | | | | | |
| Env. Economics | 0 | 0 | 1 | | | | | |
| Geography | 0 | 0 | 3 | 0 | | | | |
| Interdisciplinary | 8 | 3 | 4 | 0 | 6 | | | |
| Political ecology | 0 | 0 | 0 | 0 | 0 | 0 | | |
| Political economy | 13 | 10 | 3 | 1 | 12 | 4 | 12 | |
| Resilience | 5 | 0 | 1 | 1 | 3 | 0 | 4 | 4 |

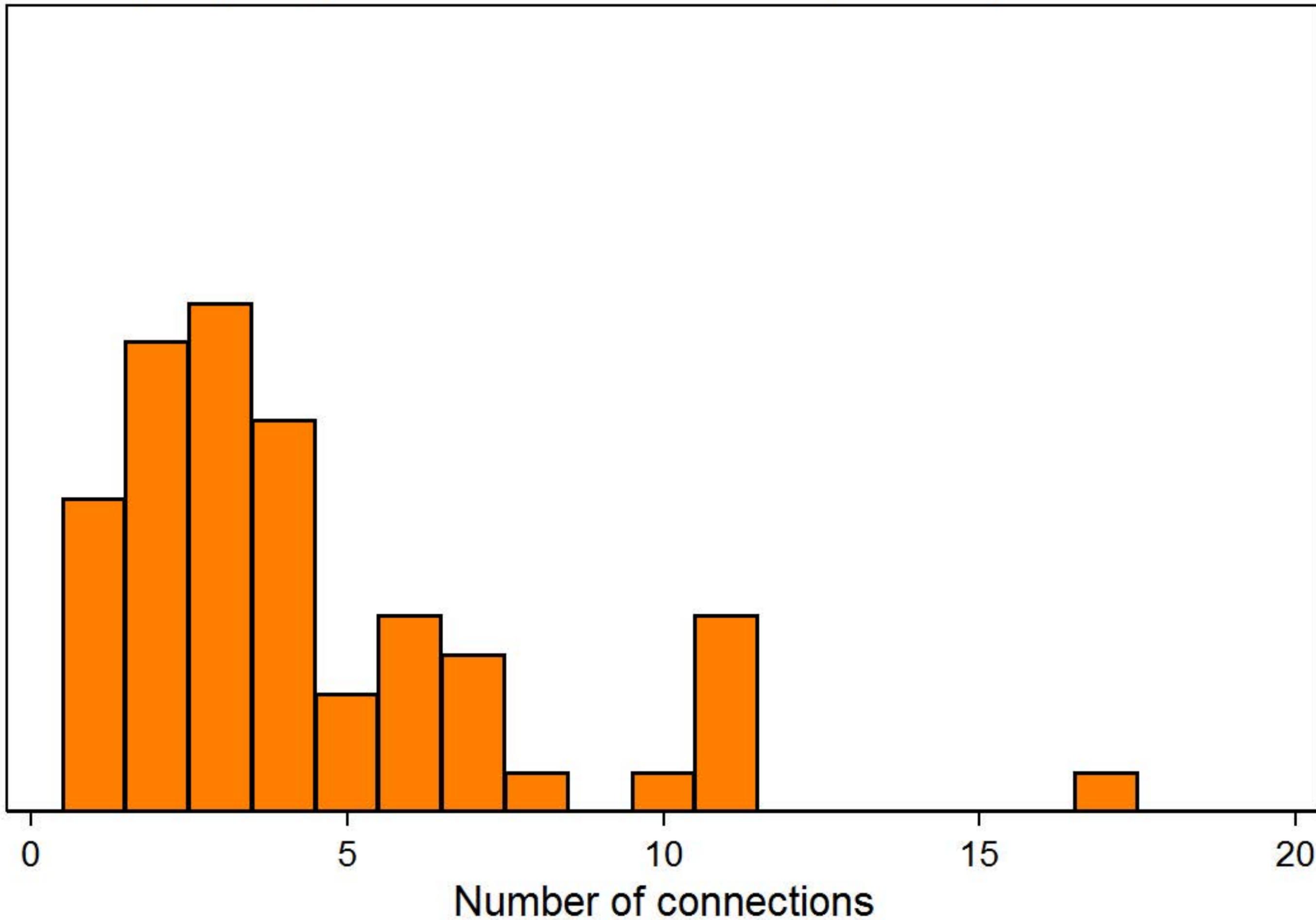
Table 6: Correlation matrix of theory attributes

| | Variables | Theories | Related | Nested in | Contains | Contradictory | Steps | List | Moderators |
|----------------------|------------------|-----------------|----------------|------------------|-----------------|----------------------|--------------|-------------|-------------------|
| Variables | 1.00 | | | | | | | | |
| Theories | 0.44 | 1.00 | | | | | | | |
| Related | 0.27 | 0.66 | 1.00 | | | | | | |
| Nested in | -0.26 | 0.28 | -0.32 | 1.00 | | | | | |
| Contains | 0.73 | 0.43 | 0.07 | -0.18 | 1.00 | | | | |
| Contradictory | 0.22 | 0.65 | 0.58 | -0.07 | 0.02 | 1.00 | | | |
| Steps | 0.36 | 0.11 | 0.07 | -0.16 | 0.29 | 0.09 | 1.00 | | |
| List | 0.87 | 0.48 | 0.21 | -0.19 | 0.85 | 0.14 | 0.30 | 1.00 | |
| Moderators | 0.31 | -0.02 | 0.06 | -0.19 | 0.04 | 0.10 | -0.40 | 0.00 | 1.00 |

(1B) Distribution of theories by the number of variables they use



(1C) Degree Distribution of Theory-Theory Network



(1A) Distribution of variables by the number of theories using them

